A Comparison of NO and N2O Production by the Autotrophic Nitrifier *Nitrosomonas europaea* and the Heterotrophic Nitrifier *Alcaligenes faecalis*

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Soil microorganisms are important sources of the nitrogen trace gases NO and N₂O for the atmosphere. Present evidence suggests that autotrophic nitrifiers such as *Nitrosomonas europaea* are the primary producers of NO and N₂O in aerobic soils, whereas denitrifiers such as *Pseudomonas* spp. or *Alcaligenes* spp. are responsible for most of the NO and N₂O emissions from anaerobic soils. It has been shown that *Alcaligenes faecalis*, a bacterium common in both soil and water, is capable of concomitant heterotrophic nitrification and denitrification. This study was undertaken to determine whether heterotrophic nitrification might be as important a source of NO and N₂O as autotrophic nitrification. We compared the responses of *N. europaea* and *A. faecalis* to changes in partial O₂ pressure (PO₂) and to the presence of typical nitrification inhibitors. Maximal production of NO and N₂O occurred at low PO₂ values in cultures of both *N. europaea* (PO₂ 0.3 kPa) and *A. faecalis* (PO₂ 2 to 4 kPa). With *N. europaea* most of the NH₄⁺ oxidized was converted to NO₂⁻, with NO and N₂O accounting for 2.6 and 1% of the end product, respectively. With *A. faecalis* maximal production of NO occurred at a PO₂ of 2 kPa, and maximal production of N₂O occurred at a PO₂ of 4 kPa. At these low PO₂ values there was net nitrite consumption. Aerobically, *A. faecalis* produced approximately the same amount of NO but 10-fold more N₂O per cell than *N. europaea* did. Typical nitrification inhibitors were far less effective for reducing emissions of NO and N₂O by *A. faecalis* than for reducing emissions of NO and N₂O by *N. europaea*. *A. faecalis* produced much less NO and N₂O under denitrifying conditions than under nitrifying conditions, and the NO produced appeared to result primarily from chemical interactions involving NO₂⁻ at pH 6.95. Once much of the nitrite was consumed, the NO and N₂O produced were further reduced to N₂. Given the rates of NO and N₂O production reported here, our results suggest that heterotrophic nitrification may be a significant source of N₂O in aerobic to near-anaerobic soils and water.

Losses of nitrogen trace gases may result in decreasing fertility, especially in those ecosystems which have suffered disturbance (30). In addition, both nitric oxide (NO) and nitrous oxide (N₂O) affect the earth’s atmosphere. In the troposphere NO controls the concentration of ozone, whereas N₂O behaves as a greenhouse gas. In the stratosphere N₂O causes destruction of ozone (18, 29). Efforts to estimate global or regional sources and sinks for NO and N₂O require an understanding of the mechanisms of the production of these compounds so that the variables controlling fluxes of these gases can be identified.

NO and N₂O are produced by a wide variety of organisms, including autotrophic nitrifiers, heterotrophic nitrifiers (including both bacteria and fungi), heterotrophic denitrifiers, dissimilatory nitrate reducers, nitrate respirers, plants, and algae. It has become clear that there is a good deal of overlap in the capabilities of the various categories of organisms. For example, current evidence suggests that many organisms, such as *Pseudomonas* spp. and *Alcaligenes faecalis*, which can denitrify under anaerobic conditions can also nitrify under aerobic conditions (5, 20, 24). The mechanisms of production of NO and N₂O by autotrophic nitrifiers such as *Nitrosomonas europaea* and heterotrophic denitrifiers such as *A. faecalis* are probably the same (i.e., reduction of NO₂⁻ to N₂O). It is not altogether clear how strict aerobic or anaerobic conditions must be to support these processes.

The results of most field studies have suggested that NO and N₂O fluxes from intact aerobic soils are primarily due to metabolism by autotrophic nitrifiers such as *N. europaea* (2, 28), although in acid forest soils heterotrophic nitrifiers may be responsible for these fluxes (19). In anaerobic soils NO and N₂O are produced by denitrifiers (1, 26). The net emission of either NO or N₂O from soils depends upon the relative importance of production and consumption processes and is, therefore, highly sensitive to physical variables, such as soil texture and water-filled pore space (9, 26). In most field studies it is not at all apparent whether autotrophic nitrifiers, denitrifiers, or perhaps denitrifiers behaving as heterotrophic nitrifiers are responsible for emissions of NO and N₂O. Probably all of these organisms play a role in producing N trace gases depending on subtle changes in soil physical and chemical conditions (6).

Heterotrophic nitrifiers have not generally been considered important sources of nitrogen trace gases in terrestrial or aquatic ecosystems, yet these organisms are widespread and may dominate nitrogen-cycling processes in some ecosystems (for example, acid forest soils) (25). *A. faecalis* is capable of producing NO, N₂O, and NO₂⁻ in medium containing both NH₄⁺ and NO₃⁻ at partial O₂ pressures (pO₂ values) ranging from 20 kPa to almost anaerobic conditions, a process that has been referred to as aerobic denitrification (1). More recently, however, this production of NO and N₂O as well as NO₂⁻ has been shown to result from oxidation of

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NH₄⁺ or organic N compounds, in which organic C is used as a carbon and energy source (5, 20). In this study we compared the abilities of the heterotrophic nitrifier A. faecalis and the autotrophic nitrifier N. europaea to produce N trace gases under various pO₂ conditions and in the presence of a variety of nitrification inhibitors.

MATERIALS AND METHODS

Culture conditions. N. europaea ATCC 19718 and A. faecalis ATCC 8750 were purchased from the American Type Culture Collection, Rockville, Md. N. europaea was maintained in a medium containing 0.5 g of (NH₄)₂SO₄, 1 g of K₂HPO₄, 0.03 g of FeSO₄, 7H₂O, 0.3 g of MgSO₄·7H₂O, and 7.5 g of CaCO₃ in 1 liter of distilled water. N. europaea was grown in a medium containing 1 g of (NH₄)₂SO₄, 0.5 g of Na₂CO₃, 0.04 g of CaCl₂, 2H₂O, 0.2 g of KH₂PO₄, 0.0005 g of ferric citrate, 0.1 ml of 0.05% phenol red solution (Flow Laboratories), and 11.9 g of HEPES buffer (N-2-hydroxyethylpiperazine-N'-2-ethanesulfonic acid) (Sigma Chemical Co., St. Louis, Mo.) in 1 liter of distilled water adjusted to pH 8.15.

A. faecalis was grown in a citrate minimal medium containing 9.2 g of K₂HPO₄ and 1.81 g of KH₂PO₄ in 1 liter of distilled water adjusted to pH 6.95 and autoclaved prior to addition of the following filter-sterilized (pore size, 0.2 μm) reagents: 661 mg of (NH₄)₂SO₄, 24.2 mg of Na₂MoO₄·2H₂O, 5.6 mg of FeSO₄·7H₂O, 0.99 mg of MnCl₂·4H₂O, 51.5 mg of CaCl₂·2H₂O, 200 mg of MgSO₄·7H₂O, and 1.18 g of sodium citrate. For denitrification studies performed with A. faecalis, KNO₃ was added to a final concentration of 0.5 to 1 mM.

Batch culture experiments. Batch culture experiments with continuous sparging were started by inoculating media (600 ml) with filtered and washed (unless otherwise stated), exponentially growing bacteria. Exponentially growing N. europaea cultures were produced by diluting aliquots of a maintenance culture 10-fold into growth medium and incubating the resulting preparations at 32°C for 2 to 3 days with shaking at 150 rpm until a cell density of 5 × 10⁹ to 10 × 10⁹ cells per ml was reached. Exponentially growing A. faecalis preparations were produced by growing cultures in citrate minimal medium inoculated from a Trypeticase soy agar slant for approximately 2 days with shaking (150 rpm) at 32°C. Experimental cultures were incubated at 32°C and continuously sparged with mixtures of filtered (pore size, 0.2 μm) air and N₂ at a flow rate of 150 ml min⁻¹, with the gas flow rates and mixing ratios controlled by mass flow controllers (Tylan Corp., Torrance, Calif.).

Dead-cell controls were produced either by holding samples of exponentially growing cultures in a boiling water bath for 30 min or by sterilizing them in a microwave oven at 700 W for 5 min (14).

Batch culture experiments in serum bottles were started by diluting exponentially growing cultures, prepared as described above, eightfold into fresh media (20 ml) in 125-ml serum bottles. The bottles were sparged with filtered air-N₂ mixtures for 30 min after the bottles were capped.

Batch culture experiments in flasks without sparging were prepared by diluting filtered and washed, exponentially growing cultures, prepared as described above, into fresh media (100 ml) in 500-ml flasks. When anaerobic conditions were required, the medium was kept in a Coy anaerobic chamber for 2 days prior to the start of the experiment. The filtered cells and all amendments were added in the anaerobic chamber before the flasks were sealed.

Determination of pO₂ in medium. A peristaltic pump was used to pump medium that was continuously sparged with air-N₂ mixtures past a Clark type dissolved oxygen (DO) microelectrode (Microelectrodes, Inc., Londonderry, N.H.) whose Teflon membrane was cemented into a 0.125-in. (0.32-cm) polyethylene tee. Prior to the start of the experiment, the tee was sterilized with ethanol (70%, vol/vol) and rinsed with sterile, distilled water. Calibration was performed before bacteria were added by bubbling the medium with N₂ for at least 2 h and setting the DO meter on zero. The DO meter was then set at full scale after the medium was sparged with either 1 kPa of O₂ in N₂ or breathing air. The sparging was continued until the reading stabilized. At the end of the experiment the DO electrode was again placed in sterile medium sparged with N₂ in order to determine whether the zero point of the DO meter had drifted. Over an 8-h period we observed drift of approximately 10% of full scale when 1 kPa of O₂ was used to set full scale and drift of 1 to 2% when 20 kPa of O₂ was used to set full scale.

Bacterial enumeration. N. europaea cultures were counted directly either in a Petroff-Hauser chamber by using phase-contrast microscopy and a magnification of ×540 or on filters (pore size, 0.2 μm; Nuclepore) by using an epifluorescence microscope after staining with acridine orange. For A. faecalis enumerations, a standard curve was constructed for optical density at 600 nm versus cell number, as determined by serial dilution and plating on Trypticase soy agar; cell counts were based on optical densities at 600 nm.

Chemical analyses. Ammonia was determined by the method of Solorzano (27). Nitrite and nitrate were analyzed as described by Parsons et al. (21). Samples taken for these determinations were filtered (pore size, 0.2 μm) and stored refrigerated in sterile vials until analyses were performed.

Samples for N₂O analysis were taken with disposable plastic syringes equipped with rubber plunger tips and three-way delrin stopcocks. Samples (10 ml) were injected within 30 min into a 1-ml gas sampling loop of a gas chromatograph (model 5890; Hewlett Packard) equipped with an electron capture detector and a packed Poropak Q column (3.5 m by 0.3 cm). The detector temperature was 330°C, and the oven temperature was 50°C. Argon containing methane (5%) was supplied as the carrier gas at a flow rate of 30 ml min⁻¹.

Samples for analysis of NO were taken with disposable plastic syringes equipped with rubber-tipped plungers and three-way delrin stopcocks. These samples were analyzed immediately. A nitrogen dioxide detector (model LMA-3; Scintron/Unisearch, Toronto, Canada) was modified to analyze the NO in syringe samples as shown in Fig. 1. The internal pump of the model LMA-3 detector was disabled to allow control of the air flow into the instrument with mass flow controllers (model FC-280; Tylan). The airstream (1,097 ml min⁻¹) was directed by a three-way stainless steel valve either through a CrO₃ converter which converts NO to N₂O with 100% efficiency or through a blank tube. The converter tube was made of opaque tubing (7.6 cm by 0.64 cm [outside diameter]) packed with 10% chromium trioxide on 30/60-mesh firebrick (Chromosorb P; Johns Manville Corp.), prepared as described by Levaggi et al. (16). The airstream was passed through a Nafion drier (Permapure, Inc., Toms River, N.J.) packed with indicating silica gel to control humidity. Calibration standards were freshly prepared before each experiment by mixing an NO standard (Scott Environmental Technology, Plumsteadville, Pa.) by
using mass flow controllers (Tylan). Standards were prepared daily and were kept in 1-liter Tedlar bags fitted with both a sampling septum and a gas inlet (Scott Environmental Technology). Volumes of gas standards (1 to 5 ml) were injected through a silicone septum into the airstream. When NO standards were stored in Tedlar bags over a 3-day period, there was approximately a 10% loss of NO in the highest concentration standard used (963 parts per billion, by volume).

The level of precision for replicate NO samples was typically 0.8%. The detection limit was 0.02 ng of N. When standards or samples containing high concentrations of NO (>3 ng of N) were injected into the instrument, we found that the luminol solution purchased from Scintrex/Unisearch became saturated, and the sensitivity of the instrument declined. A solution containing a 10-fold-higher concentration of luminol (10⁻³ M) in methanol was prepared as described by Burkhardt et al. (4) and was modified as suggested by Don Stedman (27a) for use in analyzing samples containing high concentrations of NO.

RESULTS

Previous studies on the effects of pO₂ on production of NO and N₂O by autotrophic nitrifiers have produced contradictory results (1, 17, 23). In order to resolve these differences, we utilized an experimental design in which culture medium was continuously recirculated past an oxygen microelectrode. This approach permitted more accurate determinations of pO₂ values in our experiments than were possible in previous studies. When early-stationary-phase cultures of N. europaea were diluted 10-fold into fresh medium and continuously sparged with air-N₂ mixtures having various pO₂ values, they exhibited a pO₂ optimum of 0.2 to 0.4 kPa for production of both NO and N₂O (Fig. 2). After dilution these cultures contained low concentrations of nitrite (0.2 to 0.5 mM). In contrast, when filtered and washed cells were used to inoculate similar cultures, production of NO and N₂O was undetectable, suggesting that the presence of nitrite was necessary for production of both NO and N₂O.

To determine the relative influence of pO₂ and nitrite concentrations on production of NO and N₂O, filtered and washed, exponentially growing N. europaea cells were added to serum vials containing either 1 or 20 mM nitrite. For each concentration of nitrite, one-half of the vials were sparged with 0.5 kPa of O₂, and the other half were sparged with 5 kPa of O₂. As shown in Fig. 3, production of both NO and N₂O varied as a function of NO₂⁻ concentration rather than as a function of pO₂. Dead-cell controls produced negligible amounts of both NO and N₂O.

In order to further examine the role of NO₂⁻ in the production of NO and N₂O by N. europaea, we measured production of these gases under anaerobic conditions by using various combinations of natural and artificial electron donors and acceptors. A combination of trimethylhydroquinone (TMHQ) (1 mM) as the electron donor and NO₂⁻ (1 mM) as the electron acceptor under anaerobic conditions was used to test whether NO₂⁻ could serve as an electron acceptor and source of NO or N₂O N during nitrification.

The concentrations of NO and N₂O produced when TMHQ was the electron donor were somewhat less than the concentrations produced when we used hydroxylamine (NH₂OH) (1 mM), the natural electron donor, during nitrification (Fig. 4). Cell-free controls produced negligible amounts of NO and N₂O in the presence of TMHQ and NO₂⁻.

In an attempt to determine whether NO or N₂O N is produced during oxidation of NH₄⁺ or NH₂OH rather than during reduction of NO₂⁻, we transferred filtered and washed, exponentially growing N. europaea cells to anaerobic medium containing either NH₄⁺ (1 mM) or NH₂OH (1 mM) as the electron donor and phenazinemethosulfate (PMS) (100 μM) as the electron acceptor. As expected, neither NO nor N₂O was produced when NH₄⁺ was the electron donor since NH₄⁺ oxidation is obligately aerobic.
However, when NH$_2$OH (1 mM) was the electron donor and PMS was the electron acceptor, NO was produced at a rate roughly equal to the rate observed when NO$_2^-$ was used as the electron acceptor, and N$_2$O was emitted at a much higher rate in the presence of PMS than in the presence of NO$_2^-$ (Fig. 5). Unfortunately, these results are equivocal since the product of NH$_2$OH oxidation is NO$_2^-$, which was available for reduction to NO and N$_2$O.

Although nitrification inhibitors such as N-serve and allylthiourea have been used in agriculture and are often used in field studies, the effectiveness of these compounds has been questioned (10). The intent of the experiments described below was to test the effectiveness of these inhibitors on production of NO and N$_2$O from NO$_2^-$ by *N. europaea* under reduced oxygen conditions. In the presence of the nitrification inhibitors N-serve (10 μg ml$^{-1}$ in 0.095% ethanol), allylthiourea (100 μM), and phenylacetylene (5 μg ml$^{-1}$ in 0.095% ethanol), production of NO (Fig. 6) was inhibited, and there was no net increase in NO$_2^-$ (data not shown) in cultures of *N. europaea* exponentially growing in medium containing 0.5 mM NO$_2^-$ under a headspace containing 1 kPa of O$_2$. Small amounts of N$_2$O, which may have been produced abiotically, were emitted in the presence of these inhibitors (Fig. 6).

In order to compare the capabilities of heterotrophic and autotrophic nitrifiers to produce N trace gases, the heterotrophic nitrifier *A. faecalis* was exposed to many of the same variables described above for *N. europaea*. When early-stationary-phase *A. faecalis* cells were filtered, washed, resuspended in fresh medium containing citrate (4 mM) as a carbon source and NH$_4^+$ as the sole nitrogen source, and shaken at 150 rpm, the NH$_4^+$ was oxidized to NO, N$_2$O, and NO$_2^-$ (Fig. 7). As the cells entered the stationary phase (data not shown), both NO$_2^-$ and NO were consumed, whereas N$_2$O continued to be produced at a linear rate (Fig. 7). The ratio of amount of NO produced to amount of N$_2$O produced during the linear phase of growth was 0.14 (standard deviation, 0.04). This contrasts with observations made during autotrophic nitrification, where the ratio of amount of NO produced to amount of N$_2$O produced was usually greater than 1. Under aerobic conditions *A. faecalis* was not able to utilize NO$_2^-$ as a nitrogen source and therefore was not able to grow in medium containing citrate (4 mM) and
Control flasks contained in 25-fold with PMS. NO₂⁻ (0.5 mM) was not consumed, whereas the dead-cell control did not (Fig. 9).

When an exponentially growing A. faecalis culture was continuously sparged with various mixtures of air and N₂, the optimum PO₂ for production of both NO and N₂O ranged from 2 to 4 kPa (Fig. 8). Over the range of PO₂ values studied (1 to 15 kPa), the ratio of amount of NO produced to amount of N₂O produced varied from 0.17 to 1.1, although we expected that this ratio would increase if we used higher flow rates for the sparging gas (31).

Under anaerobic conditions A. faecalis produced N₂O by denitrification of NO₃⁻ (Fig. 9) but not by denitrification of NO₂⁻ (data not shown). During denitrification N₂O was both produced and consumed. Dead cells did not consume NO₂⁻, nor did they produce any N₂O. On the other hand, NO production was as high in the dead-cell control as it was in the experimental culture in the presence of 1 mM NO₂⁻, suggesting that anaerobic production of NO resulted from chemodenitrification; however, the experimental culture consumed NO, whereas the dead-cell control did not (Fig. 9).

During heterotrophic nitrification by A. faecalis, N-serve (10 μg ml⁻¹ in 0.095% ethanol), allythiourea (100 μM), and phenylacetylene (5 μg ml⁻¹ in 0.095% ethanol) had no effect on NO production or growth (as measured by optical density at 660 nm). However, N₂O production appeared to be inhibited, especially by N-serve and phenylacetylene; N₂O production was at least partially inhibited by allythiourea (Fig. 10). N-serve and ethanol alone prevented accumulation of nitrite; however, nitrite production was only partially inhibited in the presence of phenylacetylene and allythiourea (data not shown).

A summary and comparison of the results of the experiments described above (Table 1) showed that in cultures of N. europaea continuously sparged with 5 kPa of O₂ and in cultures of A. faecalis continuously sparged at a PO₂ of 1.6 kPa, net levels of production of NO, normalized to cell number, were approximately equal; however, net emission of N₂O by the heterotrophic nitrifier was approximately 10-fold greater than net emission of N₂O by the autotrophic
nitrifier. It could not be determined from these experiments whether the differences in net emissions were due to variations in the production rates or the consumption rates for these gases. When nitrite concentrations were normalized to cell number in the *A. faecalis* cultures, they showed net consumption of NO$_3^-$ during incubation both in aerobic cultures and in anaerobic cultures. In the *N. europaea* cultures there was net accumulation of nitrite throughout the course of the experiment.

**DISCUSSION**

Estimates of nitrogen trace gas emission have been made for a variety of temperate and tropical ecosystems, ranging from seasonally dry forest to rain forest, yet it has been difficult to identify the critical environmental parameters which regulate fluxes of NO and N$_2$O over broad ranges of conditions (6, 9). Firestone and Davidson (7) have proposed a hole-in-the-pipe model suggesting that NO and N$_2$O fluxes are controlled at three levels; level one determines the relative rates of nitrification and denitrification, level two determines the relative proportions of NO and N$_2$O emitted from soils during either nitrification or denitrification, and level three determines the diffusion of these gases from the site of production to the atmosphere. The availability of O$_2$ to microorganisms is likely to be the critical parameter.
TABLE 1. Production of NO, N$_2$O, and NO$_2^-$ by autotrophic and heterotrophic nitrifier-denitrifiers

<table>
<thead>
<tr>
<th>Organism</th>
<th>Net production (ng of N day$^{-1}$ [10$^6$ cells]$^{-1}$)</th>
<th>% of NO$_2^-$ produced</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>NO</td>
<td>N$_2$O</td>
</tr>
<tr>
<td>N. europaea$^a$</td>
<td>21</td>
<td>8</td>
</tr>
<tr>
<td>A. faecalis$^b$</td>
<td>17</td>
<td>95</td>
</tr>
<tr>
<td>A. faecalis$^d$</td>
<td>0.1$^e$</td>
<td>6.7</td>
</tr>
</tbody>
</table>

$^a$ N. europaea cultures were continuously sparged with 5% O$_2$ in N$_2$ at a flow rate of 150 ml min$^{-1}$.
$^b$ A. faecalis cultures at a pO$_2$ of 1.6 kPa were continuously sparged at a flow rate of 150 ml min$^{-1}$.
$^c$ NO$_2^-$ production per 10$^6$ cells declined over the 24-h period of the experiment.
$^d$ A. faecalis cultures were grown unsparged in a flask containing N$_2$.
$^e$ NO production was by chemodenitrification.
$^f$ NO$_2^-$ was completely consumed over the 24-h period of the experiment.

affecting both level one and level two. A number of investigators have observed that pO$_2$ determines both the relative rates of nitrification and denitrification and also the ratio of amount of NO produced to amount of N$_2$O produced (1, 8, 13, 17, 23). The results of these experiments have often been contradictory; for example, Remde and Conrad (23) and Anderson and Levine (1) observed that NO emissions from cultures of nitrifiers are independent of pO$_2$, whereas Lipschultz et al. (17) noted an inverse relationship between NO emissions and pO$_2$. Our results suggest a solution to these contradictions. In sparged cultures the pO$_2$ optimum of 0.2 to 0.4 kPa for both NO and N$_2$O production observed in our experiments disappeared when NO$_2^-$ concentrations greater than 0.5 mM were added to the growth medium, suggesting that NO$_2^-$ and O$_2$ can compete for electrons removed during nitrification. Using partially purified extracts of nitrite reductase from N. europaea, Hooper (11) observed NO$_2^-$ inhibition of O$_2$ utilization during NH$_2$OH oxidation.

Both and Focht (22) were the first workers to demonstrate that NO$_2^-$ can serve as an electron acceptor during nitrification with resultant production of N$_2$O. Our results demonstrate that NO, like N$_2$O, is produced during reduction of NO$_2^-$ when TMHQ is used as an electron donor under anaerobic conditions. Using hydrazine as an electron donor, Remde and Conrad (23) similarly showed that during nitrification most of the NO produced is produced by nitrifier denitrification.

Under most field conditions, concentrations of nitrite in the soil are very low; thus, most investigators disregard nitrite as a possible controller of N trace gas emissions. When soils in arid ecosystems are first wetted, either artificially or by the first rains of the wet season, there are large but transient fluxes of NO (3). It has been suggested by Davidson (6) that nitrite, concentrated in thin water films around sites of nitrification during drying, might react biologically with organic matter to produce NO. He also pointed out that in dry savanna sites of Venezuela high NO fluxes coincided with high soil NO$_2^-$ concentrations (12). Although chemical production of NO from NO$_2^-$ at the low pH values frequently found in savanna sites is likely, an additional source of NO at these sites might be biological reduction of nitrite during nitrification.

The heterotrophic nitrifier A. faecalis ATCC 8750 described in this paper behaved somewhat differently than the (A. faecalis strain) characterized by Papen et al. (20) (strain DSM 30030). In our study NO$_3^-$ was not produced during nitrification, and although NO$_2^-$ was found in the medium,
its concentration decreased with increasing cell concentration, probably because of declining $pO_2$ values, whereas Papen et al. observed accumulation of both $NO_2^-$ and $NO_3^-$ to high concentrations. Kuenen and Robertson (15) observed that in chemostat cultures of Alcaligenes sp., $NO_3^-$ accumulation was greatest at the highest DO concentration, with little or no accumulation below 20% of air saturation.

Like Papen et al., we observed high concentrations of both $NO$ and $N_2O$ under aerobic conditions, with the ratios of amount of $NO$ produced to amount of $N_2O$ produced varying from 0.06 to 1.1 depending on the $pO_2$ in the medium and whether the culture was sparged. If NO is a precursor of $N_2O$, as is generally considered to be the case in heterotrophic denitrification, we would expect the ratio of amount of $NO$ produced to amount of $N_2O$ produced to be higher in sparged cultures and to vary directly as a function of the flow rate of the sparging gas, as was shown to be the case in Pseudomonas aeruginosa cultures (31). In sparged aerobic $A.\ faecalis$ cultures incubated with shaking, we observed a ratio of amount of $NO$ produced to amount of $N_2O$ produced of 0.07 (standard deviation, 0.06; $n = 12$), whereas in sparged, aerobic cultures the ratio of amount of $NO$ produced to amount of $N_2O$ produced varied as a function of $pO_2$ and ranged from 0.18 to 1.1. In sparged cultures Papen et al. (20) observed ratios of amount of $NO$ produced to amount of $N_2O$ produced of 0.9 in a rich peptide-meat extract medium and 0.006 in a defined medium. Whereas we observed both production and consumption of nitrite and NO, Papen et al. were not able to detect either nitrite reductase activity or nitrate reductase activity in cells from cultures which were active nitrifying $NH_4^+$. Given the extreme sensitivity of nitrite and nitrate reductases to even trace levels of $O_2$, it is undoubtedly very difficult to anaerobically wash and harvest cells in a manner that preserves activity. It is clear from both our results and those of Papen et al. that experiments such as these are best done in a chemostat in which the $pO_2$ and substrate concentrations can be controlled.

In our study, production of NO by sparged cultures of $A.\ faecalis$ under anaerobic conditions resulted from the abiotic process of chemodenitrification, although the rates of production were very low compared with the rates observed under aerobic conditions. On the other hand, NO produced under aerobic conditions appeared to be of biological origin. We based this conclusion on the fact that whereas levels of nitrite accumulation in the presence of inhibitors varied greatly, ranging from 3.2 $\mu$M with N-serve to 8.6 $\mu$M with ethanol to 29 $\mu$M with phenylacetylethanol in ethanol to 32 $\mu$M with allylthiourea to 99 $\mu$M in the absence of inhibitors, the levels of NO production were similar in all of the flasks. This conclusion needs further testing.

The responses of $N.\ europaea$ and $A.\ faecalis$ to nitrification inhibitors were quite different. In general, $N.\ europaea$ was far more sensitive than $A.\ faecalis$ was. The results of experiments in which nitrification inhibitors were used to identify the process responsible for emission of NO or $N_2O$ must therefore be interpreted with caution. More work must be done to identify inhibitors which distinguish autotrophic nitrification from heterotrophic nitrification.

In those ecosystems in which fluxes of nitrogen trace gases have been measured, there have been only rare attempts to identify the microbial process responsible for the fluxes. In most cases, it is assumed that autotrophic nitrifiers or heterotrophic denitrifiers are responsible for production of NO or $N_2O$. In forest ecosystems it has been suggested that heterotrophic nitrifiers, primarily fungi, might be responsible for much of the ammonium oxidized. Bacterial heterotrophic nitrifiers have generally been ignored, yet they are very common in most terrestrial and aquatic ecosystems. Results of Papen et al. (19) suggest that much of the NO and $N_2O$ emitted from forest ecosystems in Germany may be products of heterotrophic nitrification. The results of our comparison of NO and $N_2O$ emissions by autotrophic and heterotrophic nitrifiers described above support the hypothesis that heterotrophic nitrification may be as important a source or more important a source of NO or $N_2O$ in the atmosphere as autotrophic nitrification, at least in some ecosystems. Our ability to develop accurate assessments of total biogenic emissions will depend on our improved understanding of the processes involved.

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REFERENCES

27a. Stedman, D. Personal communication.