Corrosion of iron by sulfate-reducing bacteria
- new views of an old problem

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Abstract

About a century ago researchers first recognized a connection between the activity of environmental microorganisms and cases of anaerobic iron corrosion. Since then, such microbially influenced corrosion (MIC) has gained prominence and its technical and economic implications are now widely recognized. Under anoxic conditions (e.g. in oil and gas pipelines), sulfate-reducing bacteria (SRB) are commonly considered the main culprits of MIC. This perception largely stems from three recurrent observations. Firstly, anoxic sulfate-rich environments (e.g. anoxic seawater) are particularly corrosive. Secondly, SRB and their characteristic corrosion product iron sulfide are ubiquitously associated with anaerobic corrosion damage and, thirdly, no other physiological group produces comparably severe corrosion damage in laboratory-grown pure cultures. However, there remain many open questions as to the underlying mechanisms and their relative contribution to corrosion. On the one hand, SRB damage iron constructions indirectly through a corrosive chemical agent, hydrogen sulfide, formed by the organisms as dissimilatory product from sulfate reduction with organic compounds or hydrogen (‘chemical microbially influenced corrosion’, CMIC). On the other hand, certain SRB can also attack iron via withdrawal of electrons (‘electrical microbially influenced corrosion’, EMIC), viz. directly by metabolic coupling. Corrosion of iron by SRB is typically associated with the formation of iron sulfides (FeS) which, paradoxically, may reduce corrosion in some cases while they increase it in others. This brief review traces the historical twists in the perception of SRB-induced corrosion, considering the presently most plausible explanations as well as possible early misconceptions in the understanding of severe corrosion in anoxic, sulfate-rich environments.
When bacteria meet iron

Ever since its first production roughly 4000 years ago, iron has played a central role in human society due to its excellent mechanical properties and the abundance of its ores. Today, iron is used in much larger quantities than any other metallic material (1) and is indispensable in infrastructure, transportation, and manufacturing. A major drawback is the susceptibility of iron to corrosion. Corrosion of iron and other metals causes enormous economic damage. Across all industrial sectors the inferred costs of metal corrosion have been estimated to range between 2 and 3% of GDP in developed countries (2, 3). These costs are to a large extent caused by corrosion of iron, due to its abundant use and particular susceptibility to oxidative damage. Estimates of the costs attributable to biocorrosion of iron lack a computed basis, so that they vary widely and definite numbers cannot be given with certainty. Still, microbially influenced corrosion (MIC) probably accounts for a significant fraction of the total costs (4, 5, 6, 7) and due to its effects on important infrastructure in the energy industry (such as oil and gas pipelines), costs in the range of billions of dollars appear realistic.

Protection of iron against almost all types of corrosion can be achieved by painting or other coating. However, these measures are technically not always feasible (e.g. inside pipelines or tanks) and have a limited service life (8). Alloying of iron with more active metals such as chromium, nickel and molybdenum, on the other hand, yields stainless steels of high corrosion resistance. Still, large scale application of stainless steels is economically not achievable to any extent. As a result, corrosion-prone carbon steel (typically ≥98% Fe) is the most widely used metal in technical infrastructure such as oil and gas pipelines (2, 9).

Except for some cases caused by erosion or mechanical stress, the corrosion of iron is mostly an electrochemical process (10, 11), coupling metal oxidation to the reduction of a suitable...
oxidant. In contrast to redox reactions of non-metals, iron oxidation and reduction of the oxidizing agent must not necessarily occur at the same locality. Spatial separation of oxidative (anodic) and reductive (cathodic) reactions is possible as the metallic matrix allows the free flow of electrons from anodic to cathodic sites. Central to iron corrosion is the high tendency of the metal to give off electrons according to the anodic reaction

$$\text{Fe}^0 \leftrightarrow \text{Fe}^{2+} + 2 \text{e}^- \quad (1)$$

$$E^0 = -0.47 \text{ V}$$

(revised standard potential; 12, 13). Hydrated ferrous ions move into solution only as long as electrons, which cannot enter the aqueous phase, are removed from the surface by a suitable chemical reactant. The most common reactant in iron corrosion is molecular oxygen ($E^0' = +0.81 \text{ V}$) and corrosion of iron in oxic environments ultimately leads to the formation of various iron (hydr)oxides (‘rust’).

In the absence of oxygen, on the other hand, the most common electron acceptor for iron oxidation are protons from dissociated water. Here, the cathodic reaction is proton reduction to molecular hydrogen:

$$2 \text{e}^- + 2 \text{H}^+ \leftrightarrow \text{H}_2 \quad (2)$$

$$E^0' = -0.41 \text{ V}.$$  

Owing to the condition of electroneutrality, the anodic and cathodic half-reactions are stoichiometrically coupled, which in the case of Eqs (1) and (2) yields the net reaction

$$\text{Fe}^0 + 2 \text{H}^+ \rightarrow \text{Fe}^{2+} + \text{H}_2 \quad (3)$$

$$\Delta G^0' = -10.6 \text{ kJ (mol Fe}^0\text{)}^{-1}. $$
Ferrous iron from Eq. (3) readily precipitates in most anoxic environments (e.g. as FeCO₃) so that the activity of Fe²⁺(aq) usually remains low, thus making reaction (3) even more favorable, e.g. 

\[ \Delta G_{\text{environ}} = -27.7 \text{ kJ (mol Fe⁰)}^{-1} \text{ at } a_{\text{Fe}^{2+}} = 10^{-3} \text{ and pH } 7 \text{ (otherwise standard conditions).} \]

However, reaction (2) is ‘kinetically impeded’ (14, 15) and particularly slow at pH > 6, where proton availability is limiting (16). Hence iron corrosion is technically insignificant in the absence of oxygen or acid and iron constructions in many anoxic environments (e.g. marine sediment, water-logged soil) could, in principle, last for centuries.

However, the scenario described above changes in the presence of microorganisms, some of which dramatically accelerate corrosion kinetics. This is particularly true in environments with little or no oxygen and pH > 6, i.e. where from a purely chemical point of view, corrosion rates should be low. In technology, the phenomenon is referred to as (anaerobic) microbially influenced corrosion (MIC) or anaerobic biocorrosion. Figure 1 depicts a common example of MIC, viz. external corrosion under the disbonded coating of an iron pipeline in anoxic, sulfate-containing soil. Numerous ways by which microorganisms influence the corrosion of iron have been suggested (5, 17, 18, 19, 20). Microbial corrosion in oxic environments for instance, typically originates from localized colonization and microbial O₂ consumption at iron surfaces which can trigger preferential material loss at these sites (‘pitting’, 21, 22, 23). Additionally, dissolution of protective ‘rust’ deposits by aerobic iron-oxidizing microorganisms can influence corrosion rates (24, 25, 26). Under anoxic conditions or in systems with only temporary O₂ ingress, microbial corrosion tends to be even more pronounced. Here, corrosion results from microbial metabolic products such as organic acids (27, 28), hydrogen sulfide (12, 29, 30) or other corrosive sulfur species (31, 32, 33, 34). In addition to these indirect effects, more direct interactions between certain microorganisms and iron have been demonstrated (12, 35, 36, 37, 38, 39).
Most studies have focused on the corrosive effects of sulfate-reducing bacteria (SRB), but also other physiological groups such as thiosulfate-reducing bacteria (40), nitrate-reducing bacteria (41, 42, 43, 44), acetogenic bacteria (45) and methanogenic archaea (38, 39, 46, 47, 48, 49) have been implicated in iron corrosion. Still, the physiological group of environmental microorganisms with a suggested key role in the anaerobic corrosion of iron are the SRB (5, 6, 50, 51), which are widespread in many natural as well as engineered aquatic environments. SRB gain energy for growth by reduction of sulfate to hydrogen sulfide with electrons usually derived from the degradation of organic matter or from molecular hydrogen, which is a common fermentation product in soil, sediment and other anoxic settings (52). The suggested key function of SRB in biocorrosion is principally grounded on the following three observations. Firstly, iron in anoxic environments containing sulfate, i.e. the electron acceptor of SRB, is particularly prone to microbial corrosion (Fig. 1; 37, 53, 54). Secondly, SRB or their characteristic corrosion product FeS, are ubiquitously found on anaerobically corroded iron (54, 55, 56, 57, 58, 59). Thirdly, with corrosion rates of up to 0.9 mm Fe \(^0\) yr \(^{-1}\) (35 mpy), pure laboratory-grown SRB cultures corrode iron to an extent (12, 60, 61) that matches even severe cases of corrosion in the field. Hence, field data strongly suggest a prominent role of SRB in anaerobic iron corrosion while laboratory investigations provide the plausible mechanistic explanations (5, 12, 51).

This review will first give a brief account of the historical twists in the perception of SRB-induced corrosion. We will then comment on the question of how SRB phylogeny relates to corrosion and finally discuss in detail the respective mechanisms that are currently believed to most severely affect iron in sulfate-rich environments.

**SRB: Long-known key players in anaerobic iron corrosion**
The first evidence for an involvement of SRB in anaerobic corrosion was already provided more than a hundred years ago. In 1910, Gaines (62) reported the analysis of sulfur-rich corrosion products from anaerobically corroded iron constructions and hypothesized about a connection to the bacterial sulfur cycle. However, it was the work of von Wolzogen Kühr and van der Vlugt, which in 1934 (37) identified SRB as the prime cause of widespread iron pipe failures in the sulfate-rich soils of North Holland. These authors proposed a purely lithotrophic microbial process with iron as the only source of reducing equivalents. They attributed microbial corrosion to a prominent physiological trait, the utilization of cathodic hydrogen (Eq. 3) as sole electron donor by SRB (37, 63). The mechanistic explanation became famous as the (classical) ‘cathodic depolarization theory’. Much controversy followed in the subsequent decades. Most authors initially favored the theory (6, 64, 65, 66), while only few questioned the idea that microbial H₂ scavenging would accelerate corrosion (67, 68). With the beginning of the 1960s, the hypothesis was subjected to a series of electrochemical investigations (65, 69, 70, 71). Indeed, there seemed a connection between the ability of bacterial cultures to consume cathodic hydrogen and the stimulation (‘depolarization’) of the cathodic reaction in iron corrosion. However, despite the original lithotrophy-based concept of von Wolzogen Kühr and van der Vlugt, many of the later experiments were performed with lactate as an additional, organic electron donor for sulfate reduction. This greatly complicated the evaluation of obtained data. Costello (1974) convincingly demonstrated that hydrogen sulfide from organotrophic sulfate reduction (e.g. with lactate) was a cathodically active compound (29) and hence much of the electrochemical evidence for the ‘cathodic depolarization theory’ became disputable. The previously observed acceleration of cathodic reactions in SRB cultures (65, 69, 70, 71) could now be explained by reaction between sulfide and iron rather than by microbial consumption of cathodic H₂. Since then, occasional attempts to resurrect the theory have been made (38, 72, 73 74, 75), but to this day no culture-
based experiment has been able to demonstrate that bacterial consumption of cathodic hydrogen accelerates iron corrosion to any significant extent (39, 47, 67, 76). It should be stressed at this point that the study of a direct corrosive effect of SRB requires the use of essentially organic-free cultivation media to avoid unnecessary complication or even misinterpretation of data resulting from the corrosive effects of H₂S.

In fact it seemed at the time that much of the corrosiveness of SRB could be attributed entirely to their formation of H₂S, which is a powerful cathodic and anodic reactant (29, 65, 77). H₂S is known to rapidly react with metallic iron (net reaction: H₂S + Fe⁰ → FeS + H₂, ∆G° = −72.5 kJ (mol Fe⁰)⁻¹) thereby forming the characteristic corrosion product iron sulfide.

In the late 1960s and early 1970s, several authors demonstrated that such biogenic iron sulfides accelerated corrosion when deposited on the metal (78, 79, 80). Interestingly, sustained corrosion by iron sulfides required the presence of active populations of SRB. The exact mechanisms in this corrosion scenario have never been fully resolved (51, 81).

Furthermore, considerable attention was given to the formation of protective (‘passivating’) iron sulfide films (51, 77, 82, 83), i.e. a common phenomenon which reduces rather than accelerates the corrosion of iron by impeding the diffusion of oxidized iron (ions) from the metal surface to the bulk liquid. It is generally agreed that such thin iron sulfide layers are among the rate-controlling factors of corrosion (9, 51, 84).

Hence, until recently, SRB-induced corrosion was viewed as the result of biogenic H₂S and the catalytically active iron sulfides that are formed in the process of ‘H₂S corrosion’. In addition to these indirect effects, there remained speculation of a direct corrosiveness of SRB, often proposed in the elusive form of a ‘regeneration’ of ‘charged’ iron sulfides through microbial hydrogen consumption (55, 79, 81, 85).
In 2004, experimental evidence for a novel corrosion mechanism was furnished through isolation of SRB from enrichment cultures with metallic iron as the only electron donor (39). Apparently, sulfate reduction by these peculiar strains was directly fuelled by bacterial consumption of iron-derived electrons, without the involvement of cathodic hydrogen gas as an intermediate. In fact, while even the most efficient hydrogen-utilizing SRB did not accelerate iron corrosion as compared to sterile tests when grown in organic-free (lithotrophic) cultures, these novel isolates accelerated iron oxidation up to 71-fold under the same conditions (12). The existence of such a direct mechanism of electron uptake had previously been considered by some investigators (55, 86), but without the availability of defined model organisms for experimental validation. Recently, the process could be studied in much detail (12, 87) and the term ‘electrical microbially influenced corrosion’ (EMIC) was proposed (12). EMIC, which is fundamentally different from the corrosive effects of biogenic H$_2$S, can destroy metallic structures at rates of high technological relevance (Fig. 2; 12, 60).

While EMIC has so far only been observed in a limited number of highly corrosive SRB isolates (see next chapter), all SRB - by definition - can influence corrosion through excretion of the chemical H$_2$S (‘chemical microbially influenced corrosion’, CMIC) if sulfate and suitable electron donors are present. In conclusion, SRB act as either direct or indirect catalysts of anaerobic iron corrosion (EMIC and CMIC, respectively) and there are species-specific differences in this respect.

**Who’s who in SRB-induced corrosion? Phylogenetic distribution and ecological significance of direct corrosion by SRB**

Sulfate-reducing bacteria are found in five phylogenetic lineages, with most isolated strains being organotrophic mesophilic Deltaproteobacteria (52). Additionally, certain Archaea perform
a sulfate-reducing metabolism (88, 89) and archaeal thermophiles such as *Archaeoglobus fulgidus* may well contribute to corrosion in oil and gas producing facilities, particularly under conditions too hot to allow for the growth of their bacterial sulfidogenic counterparts (90, 91). However, there is currently only a limited number of sulfate-reducing isolates for which EMIC has been demonstrated and these are, thus far, all members of the deltaproteobacterial families *Desulfovibrionaceae* and *Desulfobulbaceae* (Fig. 3, highlighted in orange). Two of the isolates, *Desulfovibrio ferrophilus* and *Desulfopila corrodens*, have been key to the recent investigation of this new type of microbe-metal interaction (12, 39, 60, 87). Such strains have probably evaded earlier discovery as they are rapidly out-competed by ‘conventional’ organotrophic SRB in the commonly used media that employ high concentrations of organic substrates such as lactate (92, 93). It should be emphasized in this context that many of the commonly studied organotrophic SRB do not show the capability to corrode iron directly via the EMIC mechanism (12, 39, 47). Figure 3 contains a compilation of 16 sulfate reducers (highlighted in blue) that did not corrode iron when tested under lithotrophic growth conditions, i.e. in the absence of organic electron donors. Curiously, there are only few reports on organic-free enrichment cultures in the study of microbial corrosion (37, 39, 94). Still, attempts to enrich for directly corrosive SRB with iron as the only electron donor have consistently proven successful in cultures inoculated with anoxic marine sediments from a variety of geographic locations such as the North Sea, Singapore or Vietnam (39, 49, 61).

Interestingly, Dinh *et al.* (2004) and Uchiyama *et al.* (2010) isolated corrosive methanogenic archaea by omission of sulfate from otherwise similar enrichments with iron (39, 48). *Methanobacterium*-like strain IM1 (39) and *Methanococcus maripaludis* strain KA1 (48) were shown to corrode iron by direct electron uptake and the involvement of similar methanogenic strains in anaerobic biocorrosion in sulfate-limited environments seems likely. We expect the
number of sulfate-reducing and methanogenic isolates with the capability of EMIC to grow significantly if more researchers embrace the concept of lithotrophic cultivation.

The molecular mechanisms that enable certain SRB to withdraw electrons directly from iron are currently unknown. Likewise, there is presently no information whether this is a genetically fixed trait or whether also ‘conventional’ hydrogenotrophic SRB can adapt to iron utilization when exposed to it over long periods of time (12). It is assumed that direct electron uptake from iron involves outer membrane redox proteins such as $c$-type cytochromes (39), found in other microorganisms that interact with extracellular electron donors (95, 96) and acceptors (97, 98). This is certainly an exciting area for future research with possible synergies with other topics in the developing scientific discipline of ‘electro-microbiology’.

Generally, microbial uptake of electrons from extracellular surfaces is a widespread and ecologically significant process in many environments (99, 100, 101, 102). In the context of microbial corrosion it is particularly interesting to question the evolutionary roots and ecological significance of direct electron uptake. Significant quantities of anthropogenic metallic iron have only been present on Earth for approximately 4000 years and would thus represent a rather recent electron donor in the evolution of microbial physiologies. Metallic iron originating from meteorites or deep subsurfaces (103, 104), on the other hand, is very rare, although the idea of an evolution coupled to such minerals is certainly fascinating.

Still, we speculate that the EMIC mechanism may in fact represent an evolutionary undirected physiological trait. Microbes have evolved a multitude of physiological strategies to exploit other (non-$\text{Fe}^0$) solid electron donors, acceptors and electrical mediators. For example, it has previously been shown that addition of semiconductive iron minerals to syntrophic microbial cultures can accelerate the rates of substrate turnover, indicating complex bioelectric interactions.
between microbes and iron minerals (99, 100). Similarly, a recent study reports the existence of filamentous sulfate-bacteria that conduct electrons internally over centimeter distances (102) to couple distant biogeochemical processes (101). Interestingly, SRB with the ability of direct electron uptake from iron (Fe\(^0\)) appear to be abundant in nature. Several \(10^7\) cells per gram wet mass were counted in marine anoxic sediments, despite the apparent absence of man-made iron constructions (12). We hence hypothesize that the remarkable ability of certain SRB to withdraw electrons from metallic iron is in fact derived from their ability to accept electrons from other biotic and abiotic external surfaces (12).

**EMIC vs. CMIC – emerging theories in SRB-induced corrosion**

Obviously, there is no single, generalized explanation for SRB-induced corrosion. Yet a causal understanding of the basic underlying mechanisms and principles is possible. Generally, abiotic corrosion of iron in anoxic, circum-neutral environments is a very slow process if proton reduction to H\(_2\) (Eq. 2) is the most important cathodic (i.e. electron-accepting) reaction (Fig. 2D-F). However, there are particular environments that support proliferation of sulfate-reducing bacteria which, once present in sufficient numbers, can profoundly affect metal corrosion. Corrosion may proceed 70-90 times faster in the presence of SRB than under the conditions of sterile control experiments (12, 105; compare also Fig. 2A-C with Fig. 2D-F).

The aforementioned ‘classical’ cathodic depolarization theory by von Wolzogen Kühr and van der Vlugt (37) attributed such metal damage to the microbial consumption of cathodic hydrogen. In fact, this model declared microbial H\(_2\) uptake as a necessary prerequisite for hydrogen-forming corrosion (Eq. 3: \(\text{Fe}^0 + 2 \text{H}^+ \rightarrow \text{Fe}^{2+} + \text{H}_2\)) to proceed in the first place (37, 63). However, the validity of the model is questionable. Under environmental conditions, anaerobic corrosion of iron according to equation 3 is generally thermodynamically feasible, even
in the absence of anaerobic microorganisms that remove the reaction product H₂ (12, 86, 87, 105). It is rather the limiting availability of protons as well as the kinetically impeded formation of H₂ on iron that explain the low rates of anaerobic corrosion under sterile, circum-neutral conditions (e.g. as in Fig. 2D-F). In fact, experiments with hydrogen-consuming SRB and iron as the only electron donor consistently showed that SRB were capable of using cathodic hydrogen as a substrate but that this did not affect iron corrosion to any significant extent (39, 47, 67, 76, 87). In conclusion, kinetic considerations and empirical studies explicitly demonstrate that microbial consumption of H₂ cannot and does not accelerate iron corrosion.

EMIC, on the other hand, circumvents the slow, abiotic formation of cathodic hydrogen (Eq. 3) and allows SRB to utilize iron more efficiently as an electron donor by direct uptake of electrons from iron oxidation (\( \text{Fe}^0 \rightarrow \text{Fe}^{2+} + 2 \text{e}^- \)). Specialized ‘electron-consuming’ sulfate reducers can corrode iron progressively and at very high rates (Fig. 2A-C). Here, the anodic dissolution of iron results from electron consumption by sulfate reduction (Fig. 4A), i.e. a cathodic reaction that is kinetically impossible at room temperature and in the absence of biological catalysis. EMIC by SRB is characterized by the formation of large amounts of inorganic corrosion products at a distinct stoichiometry (Fig. 4A). Microbial oxidation of four moles of \( \text{Fe}^0 \) to \( \text{Fe}^{2+} \) is coupled to the reduction of only one mole of sulfate. Consequently, \( \text{H}_2\text{S} \) from this reaction precipitates one mole of FeS (which is highly insoluble) while the remaining three moles of \( \text{Fe}^{2+} \) precipitate as non-sulfidic iron minerals, e.g. with carbonate which is abundant in produced waters and other anoxic environments. Enning et al. (2012) recently calculated that less than 4% (wt/wt) of these biogenic corrosion products is actually biomass, while the remainder consists of inorganic ferrous minerals and other mineral precipitates (12). The same study (12) further demonstrated that the bulky black crusts formed through EMIC (compare Fig. 2B) are electrically conductive and hence direct contact between corrosive SRB
and the metal is not a necessary condition for corrosion. Instead, electrons flow from the corroding iron \((4 \text{Fe}^0 \rightarrow 4 \text{Fe}^{2+} + 8 \text{e}^-)\) through the electroconductive mineral crust to the crust-attached cells reducing sulfate \((8 \text{e}^- + \text{SO}_4^{2-} + 10 \text{H}^+ \rightarrow \text{H}_2\text{S} + 4 \text{H}_2\text{O})\). Venzlaff et al. (2012) performed linear sweep voltammetry on such corrosive SRB biofilms before and after inactivation of the crust-attached cells with a biocide, showing that a cathodic acceleration of iron corrosion indeed required the presence of viable ‘electron-consuming’ SRB (87). The flow of electrons from the metal to the SRB is believed to be mainly mediated by iron sulfide minerals (see reference 12 for more information) which are long known for their semiconductor properties (106, 107). Still, the discovery of a similar mechanism in certain methanogenic archaea (see previous chapter; 39, 47, 48) demonstrated that the ability to corrode iron by direct electron uptake seems to be principally independent of iron sulfides (which are not formed in methanogenic cultures). It is noteworthy that due to an imbalance between electrons entering cells and their consumption by sulfate reduction, some of the tested SRB strains that are capable of EMIC may initially even form (rather than consume) molecular hydrogen from iron in a shunt reaction (39, 61).

CMIC results from the sulfidogenic degradation of organic matter in anoxic environments (Fig. 4C). Even though most corrosion studies have focused on the effects of sulfide formed by microbial reduction of sulfate, it should be noted that also other microbial processes such as the dissimilatory reduction of thiosulfate or sulfite can produce significant amounts of corrosive sulfide (40), and hence CMIC. Intracellular oxidation of organic compounds by SRB (108, 109) is coupled to sulfide generation which, upon diffusion out of the cell, stoichiometrically reacts with metallic iron (Fig. 4B). For instance, oxidation (corrosion) of 1 g Fe\(^0\) via the CMIC mechanism requires the complete oxidation of 0.8 g acetate, a common environmental substrate of SRB (for calculation use equation C in Fig. 4).
Biogenic sulfide may initially stimulate the anodic part of the corrosion reaction by chemisorption and direct reaction with metallic iron (65, 110, 111). However, once the metallic surface is covered with inorganic corrosion products such as FeS, cathodic reactions become more important drivers of metal oxidation (82). It has been suggested that such cathodic stimulation results from biogenic dissolved sulfides (29, 84, 105, 112). Accordingly, sulfide anions act as a shuttle for bound (uncharged) protons, thereby increasing the availability of protons as electron acceptors at cathodic sites (84, 86, 113). In principal, reduction of sulfide-bound protons should occur at the metal and at FeS surfaces alike, but the latter may provide a particularly large cathodic surface area (Fig. 4B) while the former may become increasingly inaccessible due to coverage with organic and inorganic ‘biofilm’. Rapid hydrogen evolution from ‘H2S reduction’ with iron has been observed both in the presence (30) and absence (39, 84, 86) of SRB. The rates of corrosion in organotrophically grown sulfide-producing cultures vary widely, but can be high. Hubert et al. (2005) observed corrosion rates of up to 0.4 mm Fe\(^0\) yr\(^{-1}\) (16 mpy) in packed-bed bioreactors fed with a continuous inflow of lactate-containing medium (43). However, given the nature of the microbial inoculum (produced water enrichments), corrosion may not entirely have been the result of biogenic H\(_2\)S (CMIC); the presence of microorganisms capable of the EMIC mechanisms in these tests cannot be excluded. Most studies with pure laboratory-grown SRB cultures report lower corrosion rates in media with organic electron donors (6, 105, 114, 115).

SRB inflict damage on metal infrastructure in yet another way as hydrogen sulfide decelerates (poisons) the combining of hydrogen atoms into molecular hydrogen at the metal surface (116). This leads to the diffusion of a higher fraction of hydrogen atoms into the metal matrix. Combination of absorbed atomic hydrogen to hydrogen gas (H\(_2\)) within the metal - often along internal inclusions (117) - causes embrittlement of the metal (Fig. 4D). Cracking of metals under
mechanical stress (e.g. pressurized pipes) in the consequence of sulfide attack is known as sulfide stress cracking (SSC).

The role of FeS in iron corrosion

Iron sulfides (FeS) are the characteristic product of SRB-induced corrosion. They usually occur as part of a mixture of mineral and organic deposits found on anaerobically corroded iron constructions. Besides its central role in EMIC and CMIC (Fig. 4A and 4B, respectively), biogenic FeS may also, at least temporarily, lead to the protection of iron against corrosion. This is explained by the formation of tightly adherent FeS films on the metal surface, most likely by direct reaction of dissolved sulfide with metallic iron (77, 83, 111). Such films act as an effective process barrier by impeding the diffusion of ferrous ions from the metal anode to the aqueous environment (77, 118). Impediment (‘polarization’) of the anodic half reaction \( \text{Fe}^0 \rightarrow \text{Fe}^{2+} + 2 \text{e}^- \) has been frequently observed in cultures of organotrophically grown SRB (69, 71, 85). In organic-free cultures, where the predominant corrosive mechanism is EMIC, no significant slowdown of corrosion due to crust formation has been observed to date. Newman et al. (1992) stated that formation of protective FeS films occurs when dissolved sulfide concentrations exceed the concentration of dissolved ferrous ions at the unreacted metal surface, i.e. usually at high concentrations of dissolved sulfide (77). Rupture of the FeS film and local re-exposure of metallic iron results in rapid pitting corrosion (localized metal dissolution) unless further sulfide seals the exposed site. In the light of the dual role of FeS films in corrosion, it is not surprising that corrosion rates in sulfidic SRB cultures (CMIC) span at least two orders of magnitude (see reference 12 for a compilation of reported EMIC and CMIC corrosion rates). Particularly severe and progressive CMIC has been demonstrated in lactate-based media with high concentrations of
ferrous salts (119, 120); their scavenging of H$_2$S probably prevents formation of the protective FeS film and instead deposits fine suspensions of the mineral on the metal.

It has been reported that crystalline iron sulfides stimulate iron corrosion in sterile, sulfide-free incubations (78, 80, 121, 122). This is most likely due to catalysis by FeS of the chemical (abiotic) reduction of protons (Fig. 4E). The role of FeS in SRB-induced corrosion was emphasized by several authors (78, 79, 81), but the exact mechanisms are apparently complex and insufficiently understood (51). Newman et al. (1991) found the stimulatory effect of FeS to be small upon closer inspection and stated that the increased cathodic surface area provided by FeS was probably more important than its catalytic properties (82). Indeed, corrosion by FeS has so far only been reported for chemically prepared fine suspensions of the minerals (78, 79, 80, 121). Venzlaff et al. (2012) found the corrosive effect of FeS to be negligible in compact crusts formed by marine lithotrophic SRB (87). Further studies on the exact mechanisms and contribution of FeS to anaerobic corrosion are needed.

Relevance of different corrosion processes

A question of considerable technological interest is the relevance of the individual corrosion processes and their long-term rates of anaerobic metal destruction. Typical corrosion rates of unprotected steel in permanently anoxic environments range from 0.2 – 0.4 mm Fe$^0$ yr$^{-1}$ (8 – 16 mpy; ref. 12, 18; see also Fig. 1). This can in principle be explained by CMIC (Fig. 4B); corrosion rates as high as 0.4 mm Fe$^0$ yr$^{-1}$ (16 mpy) were reported in anoxic sulfidogenic cultures of SRB grown with organic substrates (43, 123). However, CMIC was usually far less pronounced in laboratory tests (6, 105, 114, 115), probably due to the formation of protective FeS deposits.
The stimulation of corrosion by iron monosulfides (Fig. 4E), on the other hand, was generally rather low ($\leq 0.06 \text{ mm Fe}^0 \text{ yr}^{-1}$, $\leq 2.5 \text{ mpy}$) with fine suspensions of the minerals (78, 80, 86) and negligible with more compact crusts (87).

Electrical microbially influenced corrosion (EMIC; Fig. 4A), on the other hand, led to progressive oxidation of metallic iron even in the complete absence of organic electron donors (Fig. 2). Corrosion rates of up to 0.9 mm Fe$^0$ yr$^{-1}$ (36 mpy) have been reported for long-term incubations during months when alkalization of cultivation media due to biocorrosion (compare Fig. 4A) was prevented by using small iron specimens in large quantities of fluid (12, 61). However, in vitro corrosion rates alone are an insufficient indicator of the relevance of the individual corrosion processes in situ. SRB capable of the EMIC mechanism corrode iron at technically highly relevant rates and would hence make for interesting targets of field surveys to better evaluate the significance of this corrosion mechanism. However, currently available strains are apparently not more closely related to each other than they are to other ‘conventional’ SRB (compare Fig. 3), so that a molecular detection of ‘EMIC SRB’ as a group based on 16S rDNA does not seem to be a promising application at this point. The analysis of corrosion products was proposed as another useful indicator (37). CMIC and EMIC produce corrosion products with inherently different relative amounts of sulfidic and non-sulfidic iron. While CMIC produces FeS as the sole mineral product, FeS accounts for only about 25% of the total iron minerals formed by the EMIC mechanism (see reference 12; compare also equations A and C in Fig. 4). This was used to quantitatively infer the contribution of EMIC to total SRB-induced corrosion from corrosion product analysis (12). Indeed it was demonstrated that serious corrosion damage of buried iron coupons in permanently anoxic marine sediment of the German North Sea was solely due to EMIC (12). More work needs to be done to better understand the relative contribution of EMIC and CMIC in different anoxic environments. Interestingly, evaluation of the original
corrosion product ratios provided by von Wolzogen Kühr and van der Vlugt (1934) (37), suggest that EMIC may have accounted for 75 to 91% of corrosion damage in their freshwater enrichments. It is tempting to speculate that early researchers already cultivated unidentified lithotrophic SRB capable of direct electron uptake from metallic iron.

Concluding remarks

This brief review discussed the microbial mechanisms that lead to progressive corrosion of iron in anoxic, sulfate-rich environments. Principally, two scenarios must be distinguished. This is firstly, the chemical microbially influenced corrosion (CMIC) of iron by hydrogen sulfide from microbial sulfate reduction with ‘natural’ organic substrates. Secondly, SRB corrode iron by direct utilization of the metal itself (Fig. 2A-C). This always occurs via direct electron uptake and only in a limited number of recently discovered SRB strains. Still, such electrical microbially influenced corrosion (EMIC) is assumed to be widespread (12, 39) and of considerable technical relevance.

CMIC and EMIC are the likely primary processes that drive iron corrosion in sulfate-containing anoxic environments. However, there are particular situations in which SRB-induced biocorrosion can be further exacerbated. Ingress of molecular oxygen (32, 33, 34, 51) into previously anoxic systems can lead to the formation of highly corrosive sulfur species from the partial oxidation of dissolved H$_2$S and biogenic FeS deposits at steel surfaces (124, 125, 126). This can even further impair metals that have already been damaged by SRB.

A better understanding of SRB-induced biocorrosion is envisioned to ultimately aid in the design of better MIC prevention and mitigation strategies for a variety of iron constructions with exposure to sulfate-containing anoxic waters, including oil and gas pipelines, ballast water tanks, steel pilings in marine applications and, more recently, offshore wind farms.
We are grateful to Friedrich Widdel for fruitful discussions. We wish to thank Rebecca Ansorge for her technical support. Funding for this work was provided through the Max Planck Society.
References


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Fig. 1. External corrosion on buried gas transmission pipeline in bog-soil of Germany.

A. Trench with coated carbon steel gas pipeline in water-logged, anoxic soil (1.4 mM sulfate; 17 mM dissolved inorganic carbon, DIC). External corrosion has occurred under disbonded coating at welding sites (arrow).

B. Welding site with corrosion pits. Disbonded asphalt coating and corrosion products (FeS / FeCO₃) were removed. Numbers indicate pit depth in millimeters. Bar, 20 cm.

C. Higher magnification of corrosion pits from a different site at the same pipeline. Bar, 2 cm.

Fig. 2. Corrosion of an iron key in the presence of Desulfovibrio ferrophilus strain IS5 (A-C) and corrosion under sterile (control) conditions (D-F). Both incubations were performed in artificial seawater medium at pH 7.3 and without addition of organic substrates (lithotrophic medium).

Electrical microbially influenced corrosion (EMIC) of key (A) led to substantial built-up of biogenic corrosion crust (B) and metal destruction (C) during 9 months. Abiotic corrosion of another key (D) in sterile medium during 27 months formed minimal corrosion product (E) and led to negligible metal loss (F) Bar, 1 cm.

A. Iron key before incubation with D. ferrophilus strain IS5.

B. Iron key with biogenic corrosion crust after 9 months of incubation with pure culture of strain IS5.

C. Residual iron after removal of the crust (B) with inactivated acid (10% hexamine in 2 M HCl) revealed 80.3% (2.7 g) iron weight loss due to corrosive activity of strain IS5. Hexamine-HCl did not dissolve Fe⁰.

D. Iron key before sterile incubation.
E. Iron key incubated in sterile artificial seawater medium. Corrosion is much less pronounced despite 27 months of incubation.

F. Residual iron after removal of corrosion products with inactivated acid (10% hexamine in 2 M HCl) revealed 2.9% (0.09 g) iron weight loss due to abiotic corrosion. Hexamine-HCl did not dissolve Fe⁰.

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**Fig. 3.** Phylogenetic tree constructed from full length 16S rRNA gene sequences of cultivated sulfate-reducing bacteria within the *Deltaproteobacteria*. The tree shows SRB isolates capable of direct electron uptake (EMIC; orange and ⚫) and hydrogenotrophic SRB that cannot corrode iron by the EMIC mechanism (blue and *). Other SRB (black) were not tested on Fe⁰. All depicted SRB corrode iron via the CMIC mechanism in the presence of suitable electron donors and sulfate. Tree does not include all cultivated SRB. I: *Desulfobulbaceae*. II: *Desulfobacteraceae*. III: *Desulfovibrionaceae*. The tree was calculated based on maximum likelihood with the ARB software package and SILVA database (127, 128). Branching with bootstrap values below 75 is not depicted. The scale bar represents 10% difference in sequence similarity. ‘Mic’-isolates are from Mori *et al.* (2010) (47). Figure adapted from Enning (2012) (61).

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**Fig. 4.** Schematic illustration of different types of iron corrosion by sulfate-reducing bacteria (SRB) at circumneutral pH. Biotic and abiotic reactions are shown. Depicted biotic reactions tend to be much faster than abiotic corrosion reactions. SRB attack iron via electrical microbially influenced corrosion (EMIC) or chemical microbially influenced corrosion (CMIC). Stoichiometry of the illustrated reactions is given in the lower panel of this figure. Please note...
that all depicted processes may occur simultaneously on corroding metal surfaces but differ in rates and relative contribution to corrosion.

A. Specially adapted lithotrophic SRB withdraw electrons from iron via electroconductive iron sulfides (EMIC). Excess of accepted electrons may be released as H₂ (via hydrogenase enzyme).

Participation of possibly buried (encrusted) SRB in sulfate reduction and hydrogen release is currently unknown.

B. Biogenic, dissolved hydrogen sulfide reacts with metallic iron.

C. Overall representation of CMIC. Organotrophic SRB produce hydrogen sulfide which reacts with metallic iron.

D. Sulfide stress cracking (SSC) of iron due to biogenic hydrogen sulfide.

E. Catalytic iron sulfides may accelerate reduction of H⁺-ions to H₂.

F. Slow, kinetically impeded reduction of H⁺-ions to H₂ at iron surfaces.

G. Consumption of H₂ from reactions E or F by SRB does not accelerate the rate of H₂ formation (no ‘cathodic depolarization’, see text).

Note that CMIC quantitatively depends on the availability of biodegradable organic matter (here schematically shown as carbon with the oxidation state of zero, CH₂O).
**Stoichiometry of corrosive reactions**

**Electrical Microbially Influenced Corrosion (EMIC):**

- **A** \( 4 \text{Fe}^0 + \text{SO}_4^{2-} + 3 \text{HCO}_3^- + 5 \text{H}^+ \rightarrow \text{FeS} + 3 \text{FeCO}_3^- + 4 \text{H}_2\text{O} \)

**Chemical Microbially Influenced Corrosion (CMIC):**

- **B** \( \text{H}_2\text{S} + \text{Fe}^0 \rightarrow \text{H}_2 + \text{FeS} \)
- **C** \( 3 \langle \text{CH}_2\text{O} \rangle + 2 \text{Fe}^0 + 2 \text{SO}_4^{2-} + \text{H}^+ \rightarrow 3 \text{HCO}_3^- + 2 \text{FeS} + 2 \text{H}_2\text{O} \)

**Sulfide Stress Cracking (SSC):**

- **D** \( 2 \text{H}_{\text{abs}} \rightarrow \text{H}_2 \)

**Abiotic:**

- **E** \( \text{Fe}^0 + 2 \text{H}^+ \xrightarrow{\text{FeS}} \text{Fe}^{2+} + \text{H}_2 \)
- **F** \( \text{Fe}^0 + 2 \text{H}^+ \rightarrow \text{Fe}^{2+} + \text{H}_2 \)
- **G** Same as **A**, but slower