

Dehalorespiration with Polychlorinated Biphenyls by an Anaerobic Ultramicrobacterium[∇]

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Anaerobic microbial dechlorination is an important step in the detoxification and elimination of polychlorinated biphenyls (PCBs), but a microorganism capable of coupling its growth to PCB dechlorination has not been isolated. Here we describe the isolation from sediment of an ultramicrobacterium, strain DF-1, which is capable of dechlorinating PCBs containing double-flanked chlorines added as single congeners or as Aroclor 1260 in contaminated soil. The isolate requires *Desulfovibrio* spp. in coculture or cell extract for growth on hydrogen and PCB in mineral medium. This is the first microorganism in pure culture demonstrated to grow by dehalorespiration with PCBs and the first isolate shown to dechlorinate weathered commercial mixtures of PCBs in historically contaminated sediments. The ability of this isolate to grow on PCBs in contaminated sediments represents a significant breakthrough for the development of in situ treatment strategies for this class of persistent organic pollutants.

Polychlorinated biphenyls (PCBs) were manufactured between 1930 and 1978, and their widespread use in high-temperature electrical coolants, hydraulic fluids, paints, carbonless paper, and as dedusting agents has resulted in their global distribution in even the most remote regions of the planet and throughout the food chain. The 2005 Priority List of Hazardous Substances (<http://www.atsdr.cdc.gov/cercla/>) published by the U.S. Agency for Toxic Substances and Disease Registry ranks PCBs fifth out of 275 substances. Ranking on this list is a combined metric based on the compound's prevalence at facilities within the United States, known or suspected toxicity, and potential for human exposure. With the discovery of *Desulfomonile tiedjei* strain DCB1 (24) in 1984, the door was opened for the study of bacteria that can reductively dechlorinate halogenated organic compounds that were manufactured for a wide range of applications throughout the 20th century. Subsequently, it was discovered that such bacteria can couple their growth to reductive dehalogenation in a process referred to as dehalorespiration (15) or halorespiration (15, 22). There has been an explosion of discoveries in this field, resulting in the identification of dozens of different species and strains that are capable of dechlorinating compounds ranging from chlorinated ethenes (19) to dioxins (5). Most of the bacteria that reductively dechlorinate toxic halogenated industrial pollutants have turned out to be members of the genus *Dehalococcoides*. Although several of these microorganisms have been successfully developed for commercially viable bioremediation of soils contaminated with chlorinated solvents, a proven effective treatment for in situ treatment of PCBs does not currently exist. As a result, the only accepted treatments for PCBs are remedial technologies such as dredging and capping, which are expensive,

disruptive to the environment, and impractical to implement over large areas and in remote locations.

Dehalococcoides ethenogenes strain 195, the first of the *Dehalococcoides* to be isolated in pure culture (19), respire with chlorinated ethenes. Since then, several other *Dehalococcoides* spp. capable of dechlorinating chlorinated ethenes have been isolated (13, 14, 28). *Dehalococcoides* sp. strain CBDB1 is capable of dehalorespiration with chlorobenzenes and chlorinated dioxins (1, 5), whereas *D. ethenogenes* strain 195 has been shown subsequently to dechlorinate chlorinated naphthalenes and a polychlorinated biphenyl (2,3,4,5,6-PCB) when grown with tetrachloroethene in sediment (11). In addition to *D. ethenogenes* strain 195, four different bacterial phylotypes are known to reductively dechlorinate PCBs (6, 10, 30). One of these, phylotype DEH10, is a member of the *Dehalococcoides* (10), but the remaining phylotypes belong to a phylogenetically diverse clade of bacteria that is distinct from the *Dehalococcoides*. Although this clade includes a number of phylotypes associated with dechlorination, most have not been cultured and isolates have not been described. One member of this group, dechlorinating strain DF-1 from Charleston Harbor (Charleston, SC), has been reported previously in a sediment-free coculture with a *Desulfovibrio* sp. (30). Reductive analysis of the 16S rRNA genes of this nonmethanogenic coculture determined that strain DF-1 is capable of dechlorinating PCBs, chlorobenzenes, and chlorinated ethenes (20, 29, 30), but for unknown reasons the microorganism could not be grown as a monoculture. Here we describe for the first time the growth of strain DF-1 by dehalorespiration of PCBs in pure culture and the reductive dechlorination of weathered commercial PCBs in soil by bioaugmentation with an isolate.

MATERIALS AND METHODS

Culture procedures. Unless stated otherwise, all bacterial culture work was done under strict anaerobic conditions in E-CI medium with 10 mM sodium formate and 173 μ M 2,3,4,5-PCB as described previously (3, 30). The PCB was added in acetone (0.1% [vol/vol] final concentration of acetone added to the culture medium). The cultures containing strain DF-1 were routinely grown in 50

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ml of medium in 160-ml serum bottles sealed with 20-mm Teflon-coated butyl stoppers (West Co., Lionville, PA). Sequential dilution series were conducted with 25 ml of medium in 60-ml serum bottles. L-Cysteine-HCl monohydrate (1.5 mM) or 0.5 mM titanium(III) nitrilotriacetate (TiNTA) was used as a chemical reductant to remove oxygen from the medium. The TiNTA was prepared as described previously (21). All cultures were incubated statically at 30°C in the dark. To test for growth effects of pH, organic buffers with a range of pK_a values were substituted for the carbonate-phosphate buffer used in E-CI medium as described previously (25).

A *Desulfovibrio* sp. isolated by Wu et al. (30) was used for preparation of cell extracts. *Desulfovibrio* isolation medium (DIM) containing 10 mM sodium lactate and 10 mM sodium sulfate (30) was used to grow the *Desulfovibrio* sp. and *Desulfovibrio vulgaris* strain Hildenborough (ATCC 29579). Both *Desulfovibrio* spp. were grown using 50 ml of DIM in 160-ml serum bottles or 10 ml of DIM in 18-ml anaerobe tubes, sealed with black butyl stoppers (Geo-Microbial Technologies, Inc., Ochelata, OK). Cultures were incubated at 30°C in the dark. *Escherichia coli* strain K-12 was grown aerobically at 37°C in the dark in Luria broth (LB) base (Difco, Detroit, MI) with 0.2% (wt/vol) glucose. Growth of the *Desulfovibrio* spp. and *E. coli* was monitored based on the optical density at 550 nm for the *Desulfovibrio* spp. in DIM and 660 nm for *E. coli* in LB medium.

Sterilized cell extracts of the *Desulfovibrio* spp., *D. vulgaris*, and *E. coli* were prepared by autoclaving 50 ml of culture (approximately 10^8 cells per ml) in a 165-ml glass serum bottle for 45 min. Extract cooled to room temperature was passed through a sterile 0.2- μ m filter in an anaerobic glove box and subsequently added to cultures (1%, vol/vol) for growth of strain DF-1.

To test for reductive dechlorination of weathered PCB, soil contaminated with 4.62 μ g/g Aroclor 1260 was collected from a drainage ditch located in Mechanicsburg, PA (40°13'54"N, 76°59'33"W). The soil sample used in the bioaugmentation experiment was untreated, black, and had a total organic content of 3,220 mg per kg soil. In total, 4 g of soil (wet weight) was inoculated (in triplicate) into 10 ml of anaerobically prepared low-saline mineral medium (26) in 25-ml anaerobe tubes sealed under N_2 - CO_2 (80:20) with Teflon septa. The medium composition was as follows (in mmol per liter): NaCl, 49.6; $MgCl_2$, 0.5; KCl, 10.2; $CaCl_2$, 1.0; NH_4Cl , 9.4; Na_2HPO_4 , 4.2; cysteine, 1.4; Na_2CO_3 , 28.3. This medium also contained vitamins and trace metals. No electron donors were added except for the components in the low-saline medium (0.0125% [wt/vol] cysteine) and residual hydrogen ($\leq 5\%$ [vol/vol]) present in the atmosphere of the anaerobic chamber used for inoculating and sampling of the microcosms. A DF1 culture grown to approximately 10^7 cells per ml with tetrachloroethene was flushed with N_2 - CO_2 to remove perchloroethene and trichloroethene prior to inoculating 2 ml into sediment microcosms. PCB analysis was performed as described by Kjellerup et al. (16). Nonbioaugmented controls included medium and soil containing indigenous microorganisms without DF1. Controls for abiotic activity included medium, soil containing indigenous microorganisms, and DF1 and were sterilized by autoclaving for 1 h on three sequential days.

PCB analysis. PCBs and chlorobenzenes from sediment-free cultures were extracted with ethyl acetate (1:5 [vol/vol, sample to solvent]) and analyzed by gas chromatography as described previously (3, 29). Weathered PCBs from contaminated sediments were extracted by sonication and analyzed as described by Dunnivant and Elzerman (7). Chlorinated ethenes were analyzed as described by Miller et al. (20). PCBs supplied to the cultures and used as GC standards were of the highest purity available (99%+) and were purchased from Accustandard (New Haven, CT). Chlorinated ethenes and chlorobenzenes of the highest purity available (98 to 99%+) were purchased from Sigma Aldrich (Milwaukee, WI).

Microscopy. Cultures of the isolated strain DF-1 were examined under phase contrast (oil immersion, 1,000 \times magnification) with a Zeiss Axiolab phase-contrast microscope (Zeiss, Thornwood, NY). In order to disrupt clusters of cells, 1-ml samples of culture were placed in a sterile 1.5-ml microcentrifuge tube and centrifuged for 10 min at 16,000 \times g. A 900- μ l aliquot of the supernatant was discarded, the pellet was resuspended in the remaining 100 μ l, and then the sample was exposed to mild sonication for 20 min (twice for a total of 40 min) in a Fisher Scientific FS20 sonicating water bath (Fisher Scientific Inc.). Staining of cells with 4',6'-diamidino-2-phenylindole (DAPI) was done as previously described (17).

Transmission electron microscopy was conducted on a Hitachi H-8000 transmission electron microscope (Tokyo, Japan) at 200 kV accelerating voltage. Samples were prepared by drying 50 μ l of the isolated strain DF-1 cells (grown in cysteine-reduced medium, centrifuged, and sonicated as described above) on a 150-mesh copper grid (EMS, Hatfield, PA) precoated with Collodion (nitrocellulose; EMS, Hatfield, PA) and a sputter-coated carbon film. Dried samples were then negatively stained with uranyl acetate (2%) and imaged after drying. Scanning electron microscopy was performed on a FEI Quanta 200 ESEM (Hillsboro, OR) at 30 kV accelerating voltage. Samples were prepared by dehydration of 100 μ l of the isolated strain DF-1 on a conductive sample stub

prepared with a sticky carbon tab (EMS, Hatfield, PA). Once dried, a Denton vacuum desk II desktop sputter coater (Moorestown, NJ) was used to deposit approximately 150 Å of gold-palladium mix onto each sample.

Competitive PCR. Dechlorinating strain DF-1 was enumerated by a competitive PCR assay (16). DNA was extracted from 1 ml of DF-1 isolate samples by using InstaGeneMatrix protocol no. 2.3. A competitor was constructed based on the primers 348F/884R by using the DNA template supplied in the competitive DNA construction kit (RR017; TaKaRa Bio Inc., Japan). Briefly, 16S rRNA gene copies per ml of isolate culture were determined according to the manufacturer's instructions (TaKaRa Bio) and by using the PCR conditions as described. PCR was conducted in 25- μ l reaction volumes using GeneAmp reagents (Applied Biosystems, CA), where the master mix contained 10 mM Tris-HCl, 75 mM KCl, 0.2 mM of each deoxynucleoside triphosphate in a mix, 1.5 mM $MgCl_2$, 1.6% dimethyl sulfoxide, 2.5 units of AmpliTaq DNA polymerase, 50 pM of each primer, and 14.75 μ l of nuclease-free water. A 0.5- μ l aliquot of DNA template and 2.5 μ l of competitor DNA in appropriate dilutions were added. For the PCR, an initial denaturation step at 95°C for 2 min was used, followed by 40 cycles of denaturation at 95°C for 45 s, primer annealing at 58°C for 45 s, and elongation at 72°C for 1 min. A final extension step at 72°C for 30 min was used, followed by a final holding step at 4°C. PCR products of the correct length were confirmed by electrophoresis using a 1.5% agarose gel. The intensity of the PCR products was measured by densitometry with the image analysis software Quantity One (Bio-Rad, Hercules, CA). One 16S rRNA gene copy per cell was assumed based on the genome sequences of *Dehalococcoides ethenogenes* (23) and strain CBDB1 (18).

DNA sequencing and analysis. The 16S rRNA gene of bacterium DF-1 was amplified from genomic DNA with primers pA and pH as described previously (8) and sequenced using the BigDye Terminator kit v3.1 (Applied Biosystems, Foster City, CA) per the manufacturer's instructions. Sequencing of purified DNA was performed on an ABI 3130 XL automated capillary DNA sequencer (Applied Biosystems, CA).

The 16S rRNA gene sequence from DF-1 and submitted gene sequences obtained from NCBI (<http://www.ncbi.nlm.nih.gov/BLAST>) were compiled and aligned using the automatic nucleic acid aligner in the BioEdit sequence alignment editor. A total of 21 sequences containing from 500 to 1,500 nucleotides were unambiguously aligned and used for calculation of trees by the neighbor-joining and FITCH approaches and using default settings in the PHYLIP software (<http://evolution.genetics.washington.edu/phylip.html>). Bootstrap analyses (1,000 replicates) were performed using the PHYLIP package.

RESULTS AND DISCUSSION

The coculture containing strain DF-1 and the *Desulfovibrio* sp. described by Wu et al. (30) was used as inoculum for the isolation of strain DF-1. Strain DF-1 did not produce colonies on solid or semisolid medium (0.5 to 2.0% Noble agar in E-CI) despite several attempts with 173 μ M 2,3,4,5-PCB, 173 μ M pentachlorobenzene, or tetrachloroethene volatilized within a sealed anaerobic glass jar used to store the plates or agar shake tubes. Isolation attempts by sequential dilution in a defined minimal medium (E-CI) (3) with a single carbon and energy source (10 mM sodium formate) and a single potential electron acceptor (173 μ M 2,3,4,5-PCB) and by adding 10 or 100 μ g/ml of ampicillin, vancomycin, neomycin, streptomycin, or chloramphenicol to the medium, which had been used successfully for isolation of *Dehalococcoides* spp. (1, 14, 19), were also unsuccessful. Increasing the concentration of vitamins and minerals by fourfold did not enhance dechlorination by the culture or enable us to isolate it by the means noted above. E-CI medium at 1 \times strength includes 1 μ g/liter of vitamin B₁₂, which has been shown to stimulate the growth and dechlorination of *Dehalococcoides ethenogenes* strain 195 (12). Finally, TiNTA was substituted as a medium reducing agent for cysteine, as the latter even in trace amounts was observed to support the growth of the *Desulfovibrio* sp. in pure culture. The first serial dilution of the culture in TiNTA-reduced medium resulted in the dechlorination of 2,3,4,5-PCB up to a 10⁻⁶ dilution. The *Desulfovibrio* sp. was no longer observed micro-

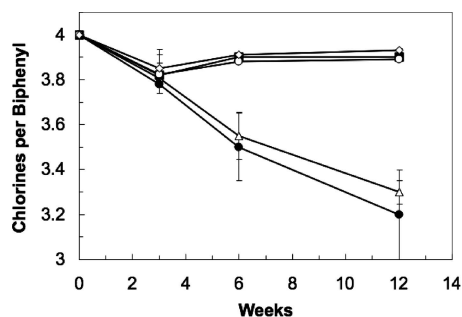


FIG. 1. Dechlorination of 2,3,4,5-PCB by strain DF-1 with sterilized culture extract from the *Desulfovibrio* sp. isolated with DF-1 (closed circles), *Desulfovibrio vulgaris* strain Hildenborough (open triangles), *E. coli* (closed squares), yeast extract (open diamonds), or with no addition (open circles). All data are the averages from triplicate cultures, and error bars represent the standard deviations. Each culture received 10 mM sodium formate and 173 μ M 2,3,4,5-PCB. Congener 2,3,5-PCB was the only dechlorination product detected.

scopically at this dilution, nor was it detected when the 10^{-6} dilution was inoculated into lactate-sulfate medium (DIM). When the 10^{-6} dilution culture was serially diluted, PCB dechlorination was observed once more to a dilution of 10^{-6} , but this time the *Desulfovibrio* sp. was not observed microscopically at any dilution, nor was growth observed from any dilution transferred to lactate-sulfate medium, LB, or LB plus glucose medium (full and half-strength), which confirmed that the culture contained only the PCB-dechlorinating strain.

When a third sequential transfer series was made in the TiNTA-reduced medium, dechlorinating activity was no longer observed, but it could be restored when the coculture of strain DF-1 and the *Desulfovibrio* sp. was reconstituted by combining a 10% (vol/vol) transfer of the isolated strain DF-1 with approximately 10^5 cells/ml of the *Desulfovibrio* sp. To determine whether active *Desulfovibrio* sp. organisms were required to restore growth of the dechlorinator, an autoclaved cell extract prepared from the *Desulfovibrio* sp. was added to the inactive strain DF-1 culture. As indicated in Fig. 1, PCB dechlorination by strain DF-1 was supported in E-CI mineral medium supplied with the *Desulfovibrio* extract, formate, and 2,3,4,5-PCB. Increasing the amount of extract added in a two- or fourfold excess had no further effect on dechlorination. An autoclaved cell extract prepared in the same manner from *Desulfovibrio vulgaris* strain Hildenborough also supported PCB dechlorination by DF-1, but autoclaved extract from yeast or *Escherichia coli* did not. No dechlorination occurred without DF-1, i.e., with cell extracts alone. Additionally, acetate (10 mM), lactate (10 mM), sulfate (10 mM), sulfide (0.1 and 1 mM), cysteine (1.5 mM), hydrogen (80:20 mix of H_2 - CO_2 at 1 atm), and DIM (1%, vol/vol) when added to E-CI medium did not support PCB dechlorination by strain DF-1 without *Desulfovibrio* extract. Strain DF-1 continued to grow to a 10^{-6} dilution after over 10 sequential transfers in medium containing *Desulfovibrio* sp. cell extract with 173 μ M 2,3,4,5-PCB or 0.2 mM tetrachloroethene. The culture was transferred back to cysteine-reduced medium without detection of the *Desulfovibrio* sp. after five sequential transfers, and contaminants were not detected based on growth in DIM, LB, or LB plus glucose and microscopic examination.

Although the contributing factor from the *Desulfovibrio* sp.

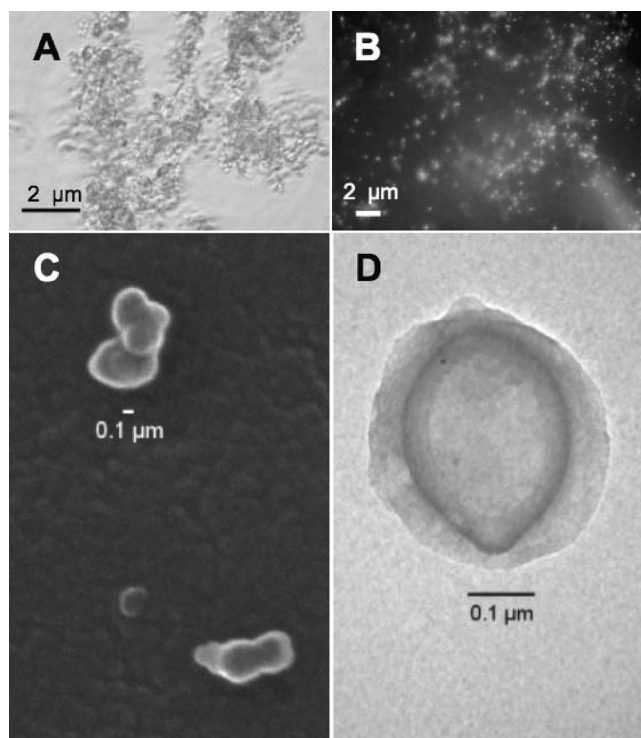


FIG. 2. Microscopic examination of dechlorinating strain DF-1. (A) Phase-contrast image before mild sonication to disrupt cell aggregates. (B) DAPI stain of DF-1 after mild sonication. (C) Scanning electron micrograph. (D) Transmission electron micrographs of negatively stained cells.

has not been identified, it appears to be specific to that genus or perhaps sulfate reducers in general. An association between microbial PCB dechlorination and sulfate reduction is not unprecedented. Several researchers have reported an inhibition or lag of PCB dechlorination by sulfate in sediments that is relieved once the sulfate has been consumed (2-4, 27). PCB dechlorination in sediment microcosms from which strain DF-1 was enriched and isolated also showed a lag in dechlorination when sulfate was present (31). Several lines of evidence support the possibility that the relationship between PCB-dechlorinating bacteria and sulfate reducers is ubiquitous in indigenous dechlorinating communities. In prior reports we discovered sulfate reducers in association with two different PCB-dechlorinating enrichment cultures (6, 30), but in neither case did the sulfate reducers prove to be directly responsible for the dechlorination. Zwiernik et al. (32) found that the addition of ferrous sulfate to sediment microcosms would stimulate PCB dechlorination after the sulfate had been consumed. Possibly, the effect that Zwiernik et al. observed was due to the delivery of a required factor from the sulfate reducers to PCB-dechlorinating bacteria. Early observations that molybdate, a specific inhibitor of sulfate reduction, also inhibited PCB dechlorination in sediment enrichment cultures (31) further suggest that sulfate reducers support the growth of PCB-dechlorinating populations.

In pure culture, strain DF-1 grew as clusters of biomass, and individual cells were rarely observed (Fig. 2A). Observation of individual cells dispersed by mild sonication and followed by staining with DAPI revealed small cocci barely visible under

TABLE 1. Reductive dechlorination of weathered Aroclor 1260-contaminated soil by bioaugmentation with bacterium DF-1^a

Congener(s)	Net reduction (ng PCB/g of soil)
PCB 63	0.439
PCB 136	0.036
PCB 82 + 151	0.083
PCB 149 + 123	0.200
PCB 153	0.250
PCB 132	0.077
PCB 163 + 138	0.339
PCB 158	0.027
PCB 183	0.088
PCB 174	0.168
PCB 202 + 171 + 156	0.077
PCB 180	0.045

^a The results show the net reduction after subtracting activity of the indigenous microbial populations in the soil. Only congeners that were reduced are shown.

the fluorescence microscope (Fig. 2B). Electron micrographs of strain DF-1 (Fig. 2C and D) showed clusters of small cocci, occasionally with slightly tapered ends. The individual cells were ultramicrobacteria that averaged 137 ± 51 nm (mean \pm standard deviation) and ranged from 75 to 339 nm in diameter ($n = 55$). This is significantly smaller than the *Dehalococcoides*, all of which have diameters in the 0.5- to 1.0- μ m range. The small size of DF-1 maximizes its surface area-to-volume ratio, which would be advantageous for a microorganism that must access a hydrophobic compound such as a PCB for its growth. Electron micrographs of the organism also revealed a structure surrounding the cells that resembled a sheath or capsule, which has not been observed with *Dehalococcoides* spp. A potentially adhesive extracellular matrix, possibly hydrophobic in nature, would be consistent with the tendency of the organisms to clump or cluster, and this may be another feature of the microorganism that facilitates its ability to absorb and accumulate hydrophobic compounds such as PCBs.

Sodium formate and hydrogen (80:20 [vol/vol] mix at 1 atm) were electron donors for PCB dechlorination by bacterium DF-1; glucose (1 mM), acetate (10 mM), lactate (10 mM), pyruvate (10 mM), propionate (10 mM), butyrate (10 mM), cysteine (1.5 mM), and sulfide (1 mM) did not support PCB dechlorination. Substitution of oxygen (air), fumarate (10 mM), nitrate (10 mM), sulfate (10 mM), sulfite (10 mM), thiosulfate (10 mM), anthraquinone-2,6-disulfonate (5 mM), ferric citrate (10 mM), and amorphous Fe(III) oxide (100 mM) for organohalides as electron acceptors with 10 mM sodium formate as electron donor did not support growth, and PCB dechlorination was inhibited when these potential electron acceptors were added with PCBs to the medium. Similar to *Dehalococcoides* spp., growth substrates for strain DF-1 are restricted to very simple electron donors (hydrogen and formate) and halogenated compounds as electron acceptors. The use of hydrogen or formate as an electron donor indicates that in sediments strain DF-1 is dependent upon a consortium of acetogens and fermenting bacteria that generate hydrogen and formate from fermentable substrates.

PCB dechlorination (2,3,4,5-PCB to 2,3,5-PCB) by strain DF-1 was maximal at 30 to 33°C, with no dechlorination observed at 10°C or 35°C after 12 weeks of incubation. Dechlorination occurred over a wide range of NaCl concentrations

(0.05 to 0.75 M) with a broad optimum (0.1 to 0.5 M) and from a pH range of 6.5 to 8.0 with an optimum at 6.8 (data not shown). Maintaining the temperature, NaCl, and pH at 30°C, 0.15 M, and 6.8, respectively, strain DF-1 in pure culture was screened for its ability to dechlorinate PCBs, chlorobenzenes, and chlorinated ethenes. The isolate reductively dechlorinated hexa- and pentachlorobenzenes, tetra- and trichloroethene, and penta- to trichlorobiphenyl congeners with double-flanked chlorines on one ring as reported previously in coculture (20, 29, 30). In the current study strain DF-1 inoculated into soil contaminated with weathered Aroclor 1260 was shown also to dechlorinate congeners ranging from octa- to pentachlorobiphenyls, further extending the range of congeners utilized by the isolate. However, the isolate consistently exhibited distinct specificity for double-flanked chlorines on one or both rings. The isolate was also tested for the ability to dechlorinate weathered Aroclor 1260 in contaminated soil (4.62 μ g/g of soil). Strain DF-1 reductively dechlorinated 8.9 mol% of congeners possessing double-flanked chlorines within 145 days (Table 1), which confirms that the strain can actively transform environmentally relevant commercial mixtures of PCBs commonly associated with impacted sites. To our knowledge this is the first demonstration of a PCB-dechlorinating isolate transforming weathered Aroclor mixtures.

Detecting growth of strain DF-1 was difficult due to its very low yields, tendency to aggregate, and small size. Determinations of absorbance, protein, dry weight, and direct counts of DAPI-stained cells did not produce reproducible values. These observations were similar to difficulties reported with some species of *Dehalococcoides* (28). Growth of strain DF-1 was successfully monitored by enumeration of 16S rRNA gene copies using a competitive PCR assay recently developed for selective monitoring of dehalogenating bacteria in sediments

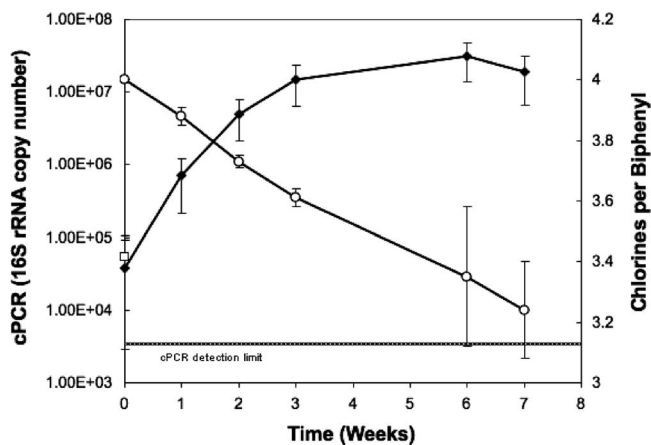


FIG. 3. Increase in 16S copy number specific to strain DF-1 in relation to dechlorination of 2,3,4,5-PCB. The open circles represent chlorines per biphenyl. The closed diamonds represent 16S copy number detected by competitive PCR from cultures grown with 2,3,4,5-PCB. The open square represents the 16S copy number detected from cultures grown without PCB at time zero (all other readings were below the detection limit of 3,500 copies). All data are the averages from triplicate cultures, and error bars represent the standard deviations. Each culture received 10 mM sodium formate and 173 μ M 2,3,4,5-PCB. Congener 2,3,5-PCB was the only dechlorination product detected.

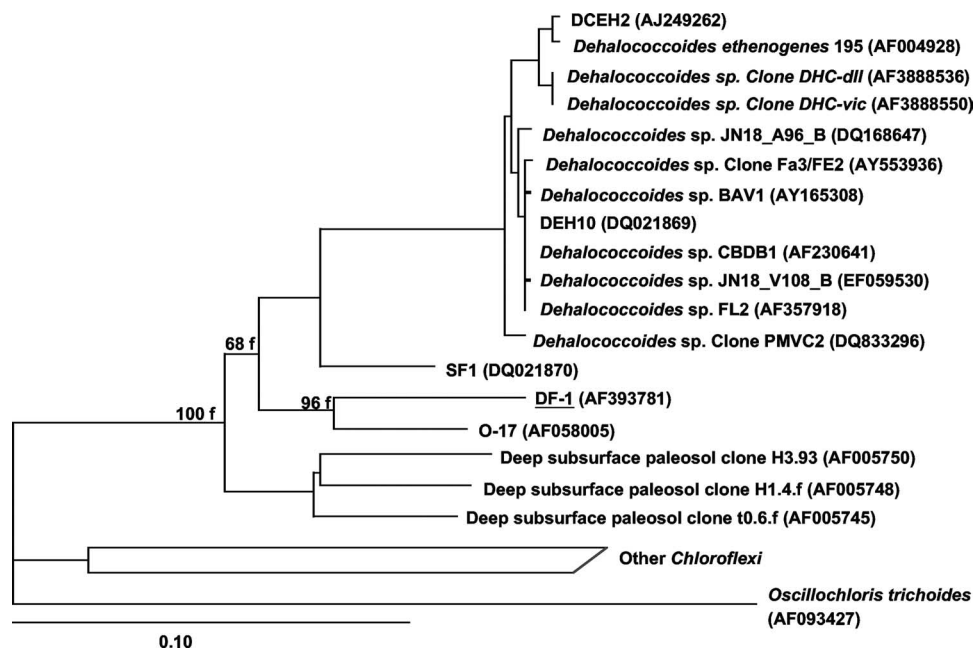


FIG. 4. Phylogenetic analysis of strain DF-1 and related 16S rRNA genes from published sequences. The tree was calculated by the neighbor-joining method and supported by FITCH. Accession numbers are shown in parentheses. Bar, 10 substitutions per 100 nucleotide positions.

(16). Using this method, DF-1 grown with 2,3,4,5-PCB showed an increase of more than 2 orders of magnitude in the copies of 16S rRNA genes, while a decrease in 16S gene copies was detected without PCBs present (Fig. 3). The data confirm that the growth of the organism is linked directly to reductive PCB dechlorination. The doubling time for DF-1 from day 0 to 14 was 2.0 days, and the growth during that period, assuming one 16S rRNA gene copy per cell, was 1.1×10^{14} cells per mol Cl^- released with the dechlorination of 2,3,4,5-PCB. Quantitative PCR methods revealed similar doubling times and cell numbers for *Dehalococcoides* sp. strains FL2 (14) and BAV1 (13) grown with chlorinated ethenes. This is the first such analysis of a PCB-dechlorinating bacterium in pure culture and the first quantification of the growth of an individual microorganism in conjunction with PCB dechlorination.

The bacterium is uniquely positioned phylogenetically based on its 16S rRNA gene sequence as the first isolate within a clade that is closely related to the *Dehalococcoides* but with less than 89% sequence similarity between the groups (Fig. 4). It most closely aligned with several phylotypes shown previously to be PCB-dechlorinating bacteria (6, 9, 10). However, this is the first member of that group to be isolated in pure culture and the first to be shown to exhibit a wide range of extracellular solute tolerances. Other distinctive phenotypic characteristics of DF-1 include its small size and synthesis of an extracellular polymer. The dechlorination of congeners with double-flanked chlorines in weathered PCBs following augmentation of contaminated sediment with strain DF-1 has significant implications for bioremediation. This capability supports the potential use of this microorganism in combination with PCB dechlorinators having complementary congener specificities (10) to stimulate the dechlorination and eventual degradation of these toxic compounds in situ.

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AUTHOR'S CORRECTION

Dehalorespiration with Polychlorinated Biphenyls by an Anaerobic Ultramicrobacterium

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Volume 74, no. 7, p. 2089–2094, 2008. Strain DF-1 was inoculated into sediments contaminated with weathered Aroclor 1260 to determine whether the augmentation would stimulate the dechlorination of congeners as they occur in the environment, adsorbed to sediment particles and in the presence of an indigenous bacterial population. The 8.9 mol% net decrease in double-flanked chlorines observed after bioaugmentation with DF-1 cannot be calculated directly from the abridged data set in Table 1 on page 2092 that highlighted only some of the changes in absolute amounts. A revised Table 1 (see following page) shows the congener profile that was used to calculate the moles percent decrease catalyzed by DF-1. There are disparities between the tables that resulted from normalization of the data in the published Table 1 to dry mass of soil. The revised moles percent analysis shows that non-double-flanked PCBs 63 and 153 did not decrease with the addition of DF-1, but a slight reduction of non-double-flanked PCBs 136 and 66/95 was significant, possibly a result of DF-1 dechlorination products serving as “primers” that stimulated the activities by the indigenous population. The relative reduction of double-flanked PCBs 180 and 202 (and coelutants) also appears to be greater. Although we cannot confirm which double-flanked dechlorination reactions were catalyzed exclusively by DF-1, the revised table clearly supports our conclusion that bioaugmentation with DF-1 stimulated reductive dechlorination of weathered Aroclor-contaminated soil.

TABLE 1. Reductive dechlorination of weathered Aroclor 1260-contaminated soil by bioaugmentation with strain DF-1

PCB(s) ^b	mol% of total PCBs in sediment ^d		
	At day 0	After 145 days	
		Without DF-1	With DF-1
15, 17	ND ^c	ND	0.04 ± 0.01
26	ND	ND	0.07 ± 0.02
51 ^d	0.08 ± 0.004	0.07 ± 0.003	0.07 ± 0.01
52, 43 ^d	0.07 ± 0.06	0.07 ± 0.04	0.11 ± 0.12
49	0.66 ± 0.29	0.04 ± 0.01	0.13 ± 0.03
47	0.18 ± 0.03	0.22 ± 0.02	0.26 ± 0.06
63 ^d	2.06 ± 1.01	2.35 ± 0.80	6.54 ± 2.48
74	0.42 ± 0.05	ND	ND
66, 95	2.57 ± 0.10	2.69 ± 0.15	2.12 ± 0.28
91	0.13 ± 0.004	0.14 ± 0.02	0.17 ± 0.08
56, 60	0.42 ± 0.005	0.42 ± 0.04	1.06 ± 0.47
92, 84, 89^d	ND	ND	6.21 ± 1.29
101	2.24 ± 0.07	2.37 ± 0.14	2.74 ± 0.48
99	0.28 ± 0.001	0.26 ± 0.03	1.48 ± 0.87
83	ND	ND	0.09 ± 0.02
97	0.09 ± 0.01	0.09 ± 0.01	0.08 ± 0.01
81, ^d 87	0.45 ± 0.03	0.48 ± 0.03	0.28 ± 0.11
136	1.66 ± 0.01	1.71 ± 0.11	1.41 ± 0.07
110, 77 ^d	1.55 ± 0.03	1.67 ± 0.11	1.93 ± 0.26
82, ^d 151	2.66 ± 0.07	2.84 ± 0.14	2.70 ± 0.05
135, 144, 124, 147^d	2.00 ± 0.02	2.08 ± 0.11	2.24 ± 0.38
149, 123^d	6.21 ± 0.16	6.74 ± 0.27	6.25 ± 0.07
118	0.41 ± 0.02	0.38 ± 0.10	0.46 ± 0.08
134	0.32 ± 0.01	0.35 ± 0.04	0.64 ± 0.20
114, 131	0.08 ± 0.02	ND	ND
146	1.40 ± 0.01	1.47 ± 0.06	3.72 ± 1.84
153	6.96 ± 0.16	7.67 ± 0.54	7.56 ± 0.22
132	2.64 ± 0.04	2.81 ± 0.14	2.60 ± 0.12
141	1.72 ± 0.01	1.82 ± 0.07	1.79 ± 0.03
137, 176, 130	0.76 ± 0.07	0.76 ± 0.02	0.83 ± 0.02
163, 138	10.50 ± 0.29	11.82 ± 1.13	9.98 ± 2.11
158	1.12 ± 0.02	1.25 ± 0.13	0.87 ± 0.42
178, 129	1.43 ± 0.02	1.56 ± 0.09	1.81 ± 0.13
175	5.36 ± 1.26	1.86 ± 2.66	1.42 ± 1.51
186, ^d 182^d	3.74 ± 0.06	4.23 ± 0.42	5.06 ± 0.29
183	2.76 ± 0.05	3.10 ± 0.29	2.75 ± 0.65
128	0.53 ± 0.003	0.60 ± 0.08	0.46 ± 0.18
185	0.48 ± 0.01	0.54 ± 0.06	0.43 ± 0.17
174	4.30 ± 0.05	4.94 ± 0.63	4.73 ± 0.93
177	2.86 ± 0.03	3.23 ± 0.41	3.80 ± 0.17
202, 171, 156	2.59 ± 0.02	1.98 ± 1.33	0.47 ± 0.03
157, 200	0.38 ± 0.01	0.43 ± 0.04	0.45 ± 0.05
172	1.34 ± 0.02	1.24 ± 0.40	ND
180	9.77 ± 0.30	9.35 ± 2.62	4.88 ± 1.48
193	0.79 ± 0.01	0.71 ± 0.22	0.49 ± 0.004
191	0.32 ± 0.01	0.31 ± 0.10	0.17 ± 0.06
199	0.29 ± 0.01	0.27 ± 0.09	0.17 ± 0.01
170, 190	4.61 ± 0.10	4.36 ± 1.27	2.06 ± 0.38
198	0.09 ± 0.002	0.09 ± 0.03	ND
201	2.65 ± 0.04	2.48 ± 0.78	1.69 ± 0.04
203, 196	3.60 ± 0.06	3.41 ± 1.01	2.12 ± 0.30
208, 195	0.89 ± 0.01	0.85 ± 0.24	0.57 ± 0.06
207	0.04 ± 0.000	0.04 ± 0.01	0.03 ± 0.001
194	1.71 ± 0.04	1.65 ± 0.45	1.13 ± 0.12
205	0.12 ± 0.005	0.12 ± 0.03	ND
206	0.34 ± 0.02	0.32 ± 0.10	0.24 ± 0.01

^aAll data are means of triplicate cultures ± standard deviations.

^bListed in order of elution. Congeners with double-flanked chlorines are shown in bold.

^cValues less than 0.01 mol% are reported as ND.

^dNot typically detected in Aroclor 1260.